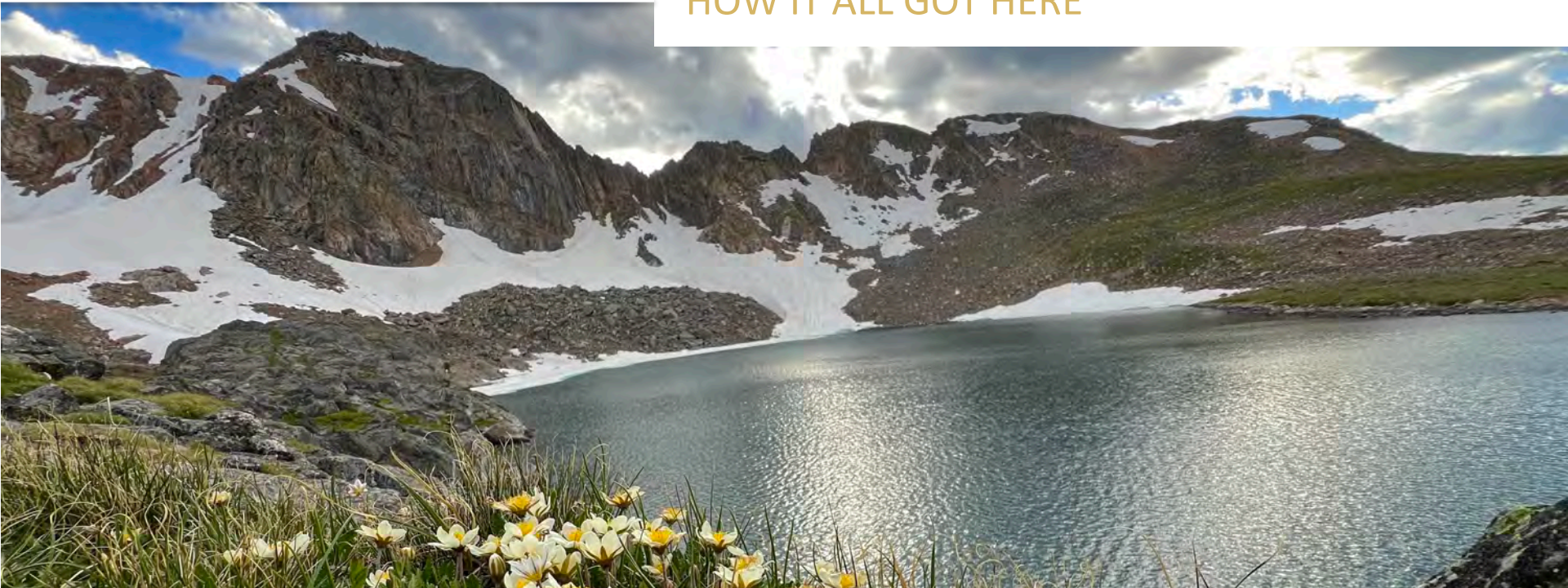




HOW IT ALL GOT HERE



WHAT

TAXONOMY & SYSTEMATICS

WHERE

SPATIAL DATA

WHEN
(and what)

GENETICS

WHY

ENVIRONMENT



Crisp Lab



Cook Lab



"Using phylogenies to test hypotheses about date, rate and mode of diversification and evolution of Australian plants and animals"



nature

Vol 468/4 April 2010/doi:10.1038/nature09764

LETTERS

Phylogenetic biome conservatism on a global scale

Michael D. Crisp¹, Mary T. K. Arroyo², Lyn G. Cook³, Maria A. Gandolfo⁴, Gregory J. Jordan⁵, Matt S. McGlone⁶, Peter H. Weston⁷, Mark Westoby⁸, Peter Willf⁹ & H. Peter Linder¹⁰

How and why organisms are distributed as they are has long intrigued evolutionary biologists^{1,2}. The tendency for species to retain their ancestral ecology has been demonstrated in distributions on local and regional scales^{3,4}, but the extent of ecological conservatism over tens of millions of years and across continents has not been assessed^{5,6}. Here we show that biome status at speciation has outweighed biome shifts by a ratio of more than 25:1, by inferring ancestral biomes for an ecologically diverse sample of more than 11,000 plant species from around the Southern Hemisphere. Status was also prevalent in transoceanic colonizations. Availability of a suitable biome could have substantially influenced which lineages establish on more than one landmass, in addition to the influence of the rarity of the dispersal event themselves. Conversely, the taxonomic composition of biomes has probably been strongly influenced by the rarity of species' transitions between biomes. This study has implications for the future because if clades have inherently limited capacity to shift biomes⁷, then their evolutionary potential could be strongly compromised by biome contraction as climate changes.

In recent years, the extent to which species retain ancestral ecological traits and environmental distributions ('niche conservatism') has been debated^{8–10}. Whether ecology is evolutionarily labile or conserved appears to be scale dependent, but most of the evidence for conservatism has come from studies at smaller (community-level) scales^{11,12}. Large-scale ecological conservatism is not well studied but there is evidence that ecological tolerances (for example as reflected in latitudinal range) are relatively conserved in lineages over long periods of macro-evolutionary time^{13–15}. This leads to the prediction that lineages might rarely colonize new biomes¹⁶; however, the frequency of transitions has not been estimated¹⁷.

Biomes are broad vegetation types defined by climate, life-form and ecophysiology¹⁸ and, hence, are useful units for investigating the large-scale pattern of ecological conservatism. Biome conservatism might be a major determinant of the global distribution of biodiversity, such as the latitudinal gradient, which could have arisen because relatively few anciently tropical taxa were able to colonize temperate biomes¹⁹. Global comparisons across multiple lineages are needed to investigate the generality of conservatism and how much it has influenced the assembly of biomes²⁰. Here we quantify the frequency of biome changes using a large sample of plant phylogenies distributed across multiple biomes and landmasses.

The widely expanding landmasses in the Southern Hemisphere provide an outstanding opportunity to assess the relative frequency of phylogenetic biome conservatism on a broad scale and over long time periods (tens of millions of years). First, each southern continent contains a rich array of biomes among which taxa could have moved. Second, many taxa are spread across the southern landmasses and are

largely restricted to that region, defining a biogeographically meaningful study area²¹. Although some of these transoceanic dispersals are likely vicariant remnants of once-continuous distributions on Gondwana^{22,23}, many more are probably the result of long-distance colonization²⁴. Third, the oceanic gaps in the Southern Hemisphere are so wide that distinct sister taxa are likely to have remained effectively separate in their new biomes, without engaging gene flow complicating comparisons of their subsequent evolution. Last, the number of plant phylogenies that are dated and well sampled at species level for the Southern Hemisphere has grown to a statistically meaningful size.

We assembled a dataset comprising 11,064 species of vascular plants in 43 taxa from subtropical Africa, Madagascar, Australia–New Guinea, New Zealand and South America, equivalent to an estimated 13% of the total flora of these regions. We reconstructed ancestral distributions and, by dating diversifications, inferred which dispersals were likely to be the result of long-distance dispersal and establishment (colonizations) across oceans. Each plant species was assigned to one or more of seven biomes (tropical forest, sclerophyll, temperate forest, temperate grassland, wetland and arid) and transitions among biomes were recorded.

We found strong support for the hypothesis of phylogenetic biome conservatism because biome shifts were rare, being associated with only 396 (3.6%) of the approximately 11,000 evolutionary diversifications. Of these shifts, 354 occurred within landmasses (Fig. 1 and 4b) were linked to transoceanic colonizations (Fig. 2 and Table 1). Investigating taxa individually, we found a consistent pattern in which closely related species were more ecologically similar than would be expected by chance (Supplementary Results), irrespective of whether the taxa have been considered Gondwanan relicts (for example *Acacia* spp., *Callitriche*, *Leptosiphon*) or otherwise (for example *Eucalyptus*, grasses and some legume tribes^{25,26}).

Most transoceanic colonizations involved no change of biome (Fig. 2a; $n = 236$ biome shifts, 40% $P = 0.0001$). Moreover, a bias towards biome status in colonizations was consistent across all landmass pairs, in both directions (Fig. 2b; Wilcoxon matched-pairs signed-rank test: $n = 30$, $z = -3.52$, $P = 0.0019$). Similarly, the bias towards status in colonizations was consistent among biomes (Table 1), whether comparing source biomes (McNemar matched-pairs signed-rank test: $n = 7$, $z = -2.15$, $P = 0.032$) or destination biomes ($n = 7$, $z = -2.19$, $P = 0.034$). We suggest that constraint on biome shifts following colonization from one landmass to another is a consequence of the dynamics of colonization. The frequency at which propagules of a species have crossed the southern oceans is likely to be extremely small. For successful colonizations, immigrant propagules must germinate and establish reproductively successful populations. Establishment should be easier in a biome to which the species is pre-adapted, that is, one approximately equivalent to its

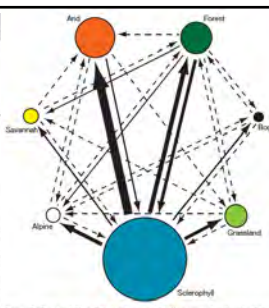


Figure 1 | Biome shifts within landmasses. Shifts occurred with only 356 of 11,064 speciation events within landmasses. Number of species sampled within each biome is proportional to the area of each circle: sclerophyll, 250 and 1,000; temperate grassland, 504; savanna, 242; temperate forest, 186; bog, 84. Arrow thickness is proportional to the number of transitions in each direction, ranging from 6 to 95 events; dashed lines indicate 1–5 events and lack of an arrow indicates that there was no event.



Figure 2 | Biome conservatism in transoceanic plant colonizations around the Southern Hemisphere. Proportions of colonizations into the source (blue) or a different (red) biome are shown along each arrow. Arrow thickness is proportional to the number of colonizations, ranging from 6 to 187 events. Dashed lines indicate fewer than 6 events. Four well-sampled colonizations are not shown.



ARTICLE

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Flammable biomes dominated by eucalypts originated at the Cretaceous–Palaeogene boundary

Michael D. Crisp¹, Geoffrey E. Burrows¹, Lyn G. Cook³, Andrew H. Thornhill¹ & David M. J. S. Bowman¹

Fire is a major modifier of communities, but the evolutionary origins of its prevalent role in shaping current biomes are uncertain. Australia is among the most fire-prone continents, with most of the landmass occupied by the fire-dependent sclerophyll and savanna biomes, in contrast to biomes with similar climates in other continents. Australia has a tree flora dominated by a single genus, *Eucalyptus*, and related Myrtaceae. A unique mechanism in Myrtaceae for enduring and recovering from fire damage likely resulted in this dominance. Here, we find a conserved phylogenetic relationship between post-fire resprouting (epicormic) anatomy and biome evolution, dating from 60 to 62 Ma, in the earliest Palaeogene. Thus, fire-dependent communities likely existed 50 million years earlier than previously thought. We predict that epicormic resprouting could make eucalypt forests and woodlands an excellent long-term carbon bank for reducing atmospheric CO₂ compared with biomes with similar fire regimes in other continents.

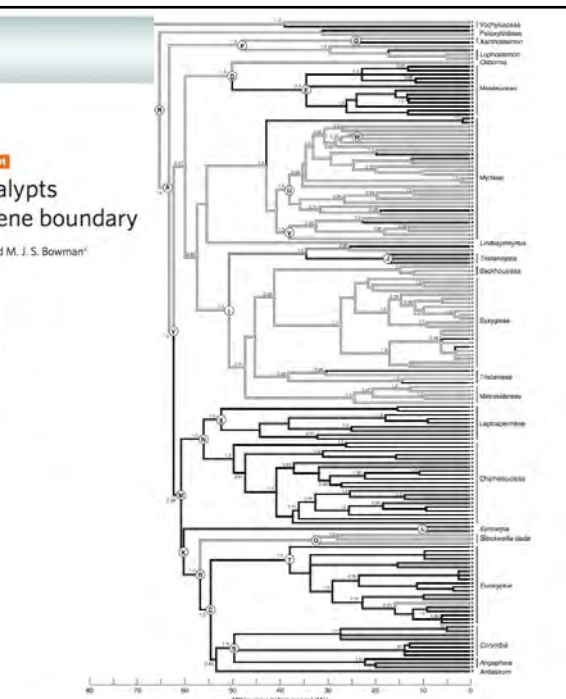
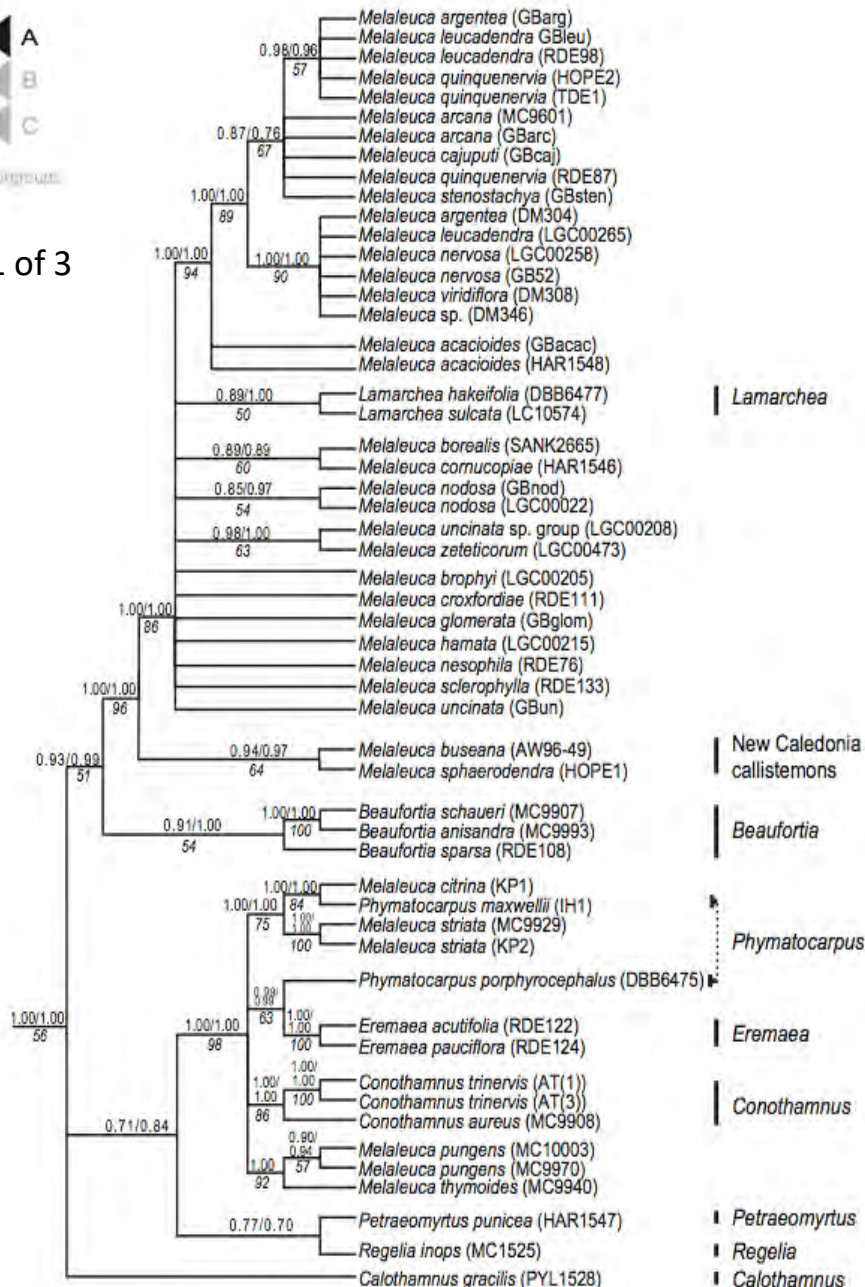
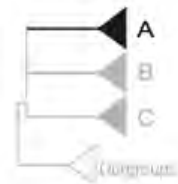


Figure 2 | Inferred evolutionary history of biome flammability mapped onto a Bayesian phylogeny of Myrtaceae. Time scale is millions of years before present (Ma). Labels indicate higher taxa mentioned in the text. Shading of boxes at tips and along branches indicates scoring for the trait 'biome flammability' (black) or 'non-flammable' (white). Ancestral states reconstructed by parsimony are shown at internal branches; grey indicates an equivocal reconstruction. Nodes labeled with upper case letters in circles are for reference from the text and tables. Decimal values on branches indicate Bayesian posterior probabilities; integers on *Corymbia* indicate parsimony bootstrap scores.

Systematics of Melaleuca



Melaleuca revisited: cpDNA and morphological data confirm that *Melaleuca* L. (Myrtaceae) is not monophyletic

Robert D. Edwards,^{1,2} Lyn A. Craven,³ Michael D. Crisp¹ & Lyn G. Cook^{1,2}

¹ The Australian National University, Research School of Biology, Canberra ACT 0200, Australia

² The University of Queensland, School of Biological Sciences, Brisbane QLD, 4072, Australia

³ Australian National Herbarium, CSIRO Plant Industry, Canberra ACT 2601, Australia

Author for correspondence: Robert D. Edwards, r.edwards2@uq.edu.au

New combinations and names in *Melaleuca* (Myrtaceae)

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¹ Australian National Herbarium, CSIRO Plant Industry, Canberra ACT 2601, Australia

² Research School of Biology, The Australian National University, Canberra ACT 0200, Australia

³ School of Biological Sciences, The University of Queensland, Brisbane QLD 4072, Australia

Author for correspondence: Lyn A. Craven, lyn.craven@csiro.au

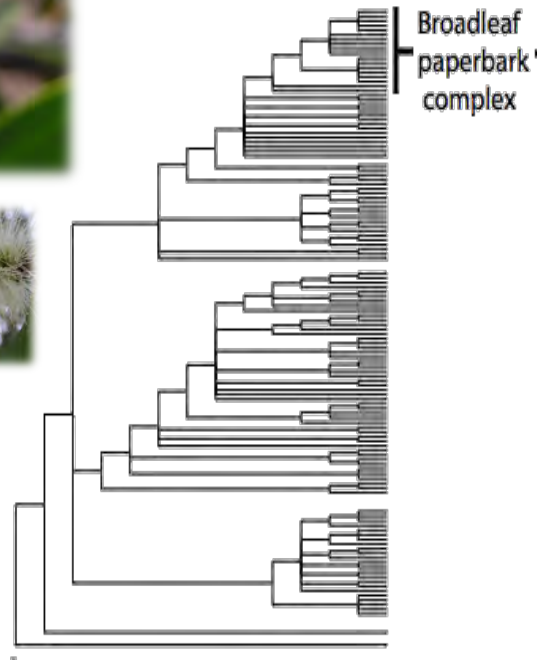
DOI: <http://dx.doi.org/10.12705/633.38>

Abstract The taxonomy of Melaleuceae has been re-assessed in the light of DNA sequence studies and it is concluded that all genera should be placed within *Melaleuca*. *Beaufortia*, *Calothamnus*, *Conothamnus*, *Eremaea*, *Lamarchea*, *Petraeomyrtus*, *Phymatocarpus* and *Regelia* and their constituent species are formally transferred to *Melaleuca*. One hundred and six nomenclatural changes are provided of which 72 are new combinations and 34 are new names.

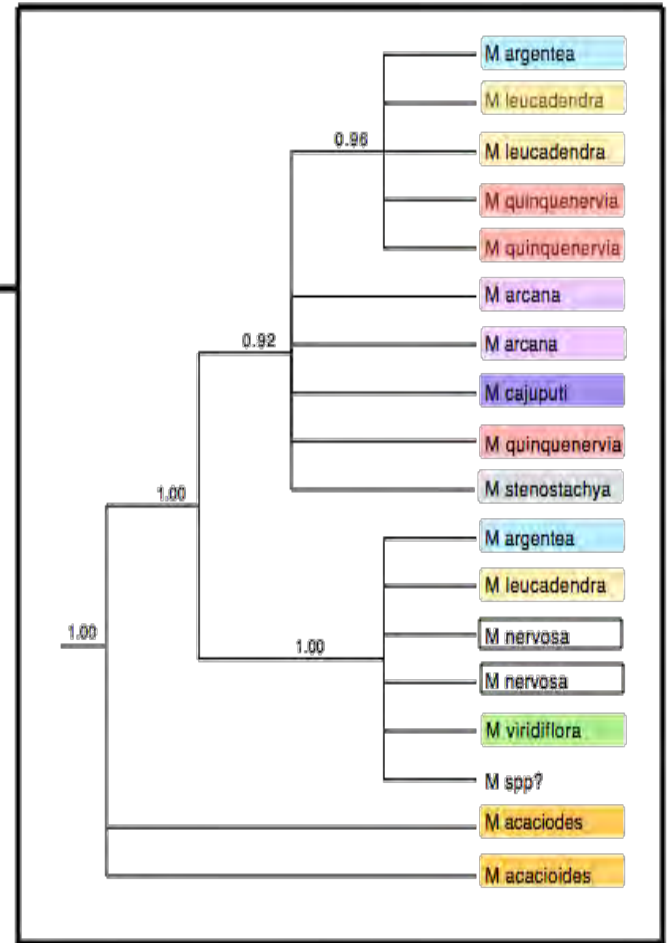
Keywords Australia; *Beaufortia*; *Calothamnus*; *Conothamnus*; *Eremaea*; *Lamarchea*; *Melaleuca*; *Petraeomyrtus*; *Phymatocarpus*; *Regelia*



Systematics of Melaleuca - Broadleaved Paperbarks



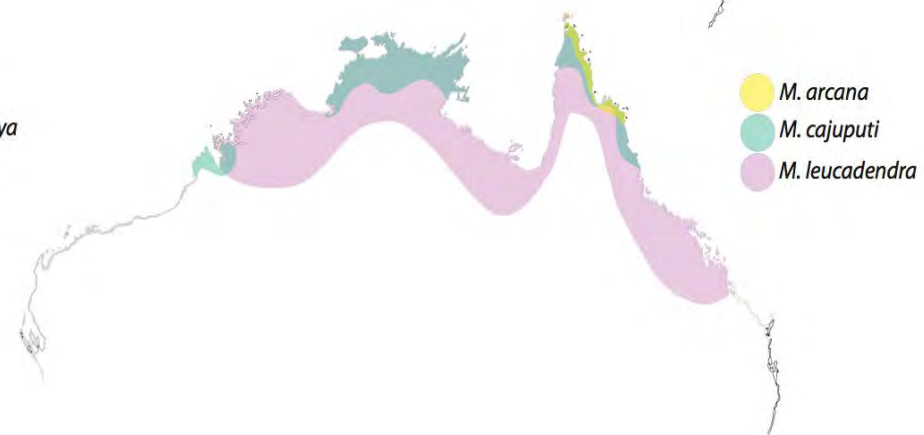
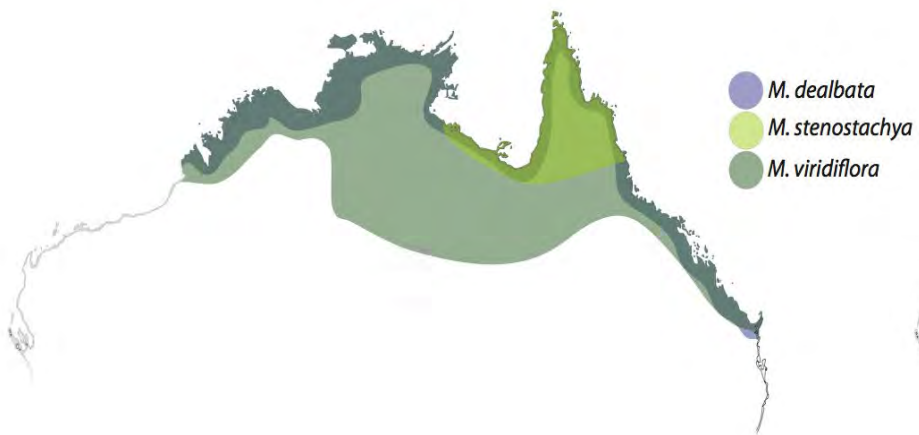
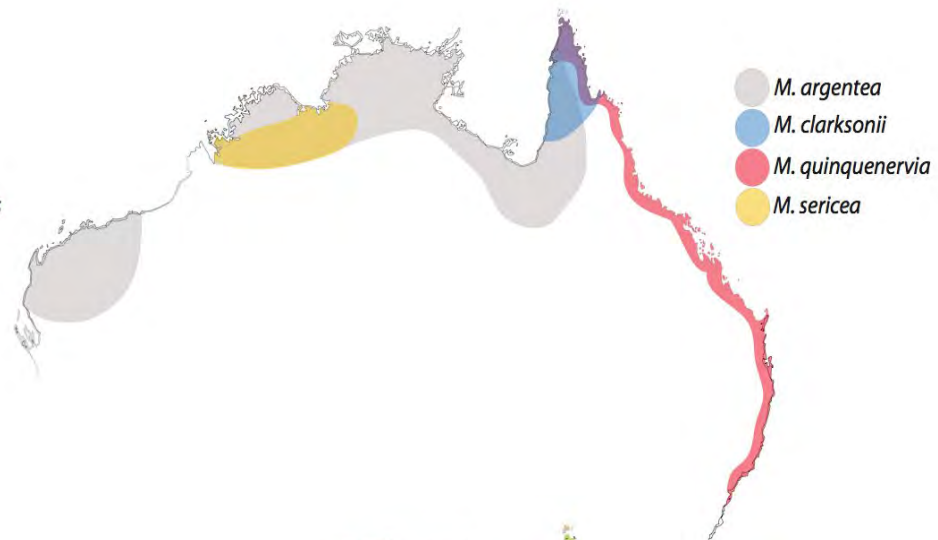
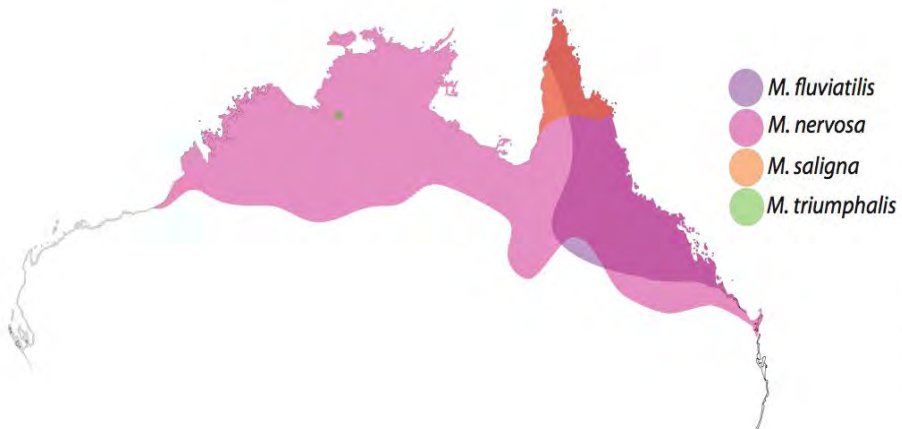
(Edwards et al., 2010)



(Cook et al., 2008)



Systematics of Melaleuca - Broadleaved Paperbarks





Systematics of Melaleuca - Broadleaved Paperbarks

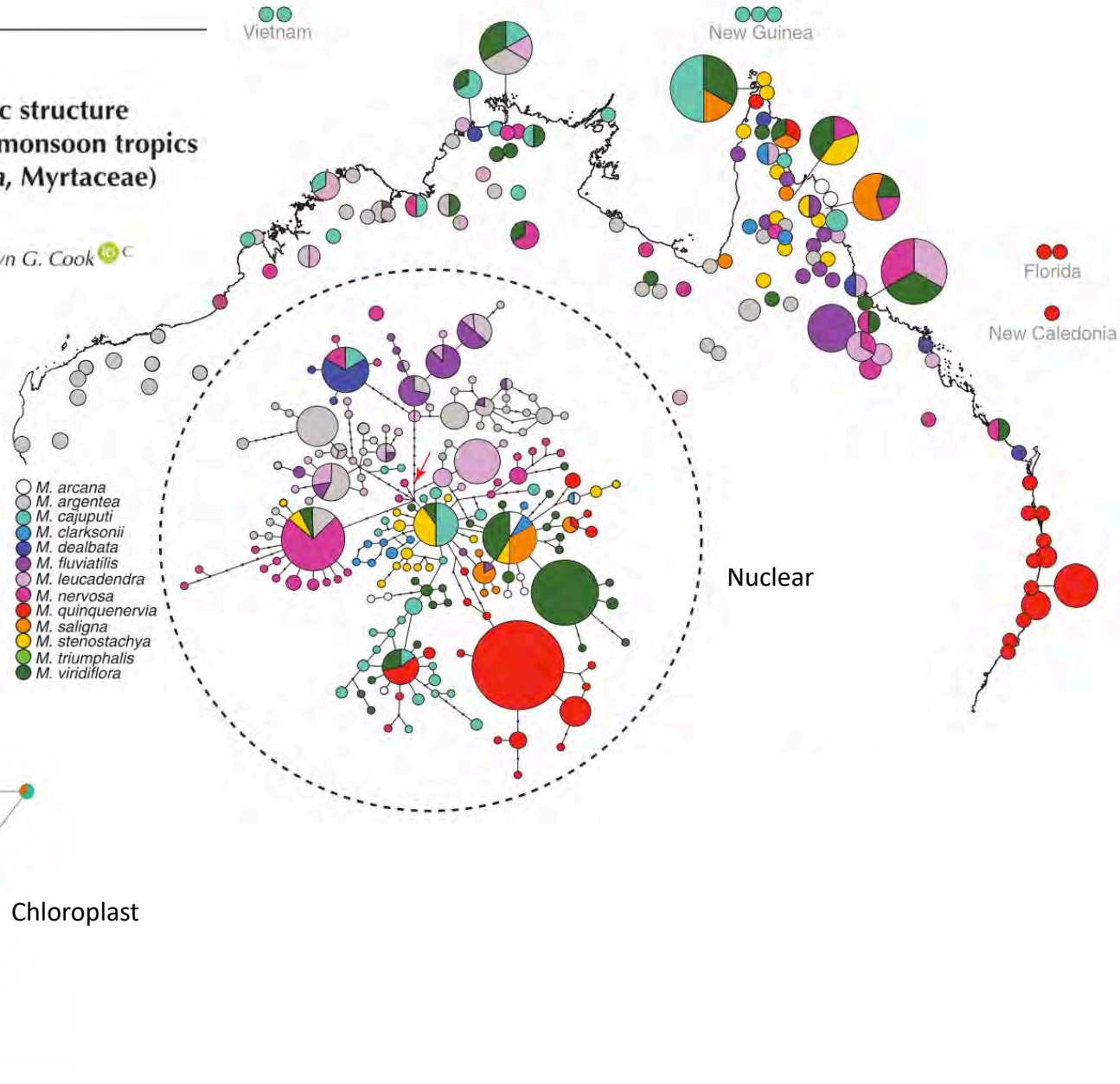
CSIRO PUBLISHING

Australian Systematic Botany, 2018, 31, 495–503

<https://doi.org/10.1071/SB18032>

Species limits and cryptic biogeographic structure in a widespread complex of Australian monsoon tropics trees (broad-leaf paperbarks: *Melaleuca*, Myrtaceae)

Robert D. Edwards^{A,D}, Michael D. Crisp^B and Lyn G. Cook^C





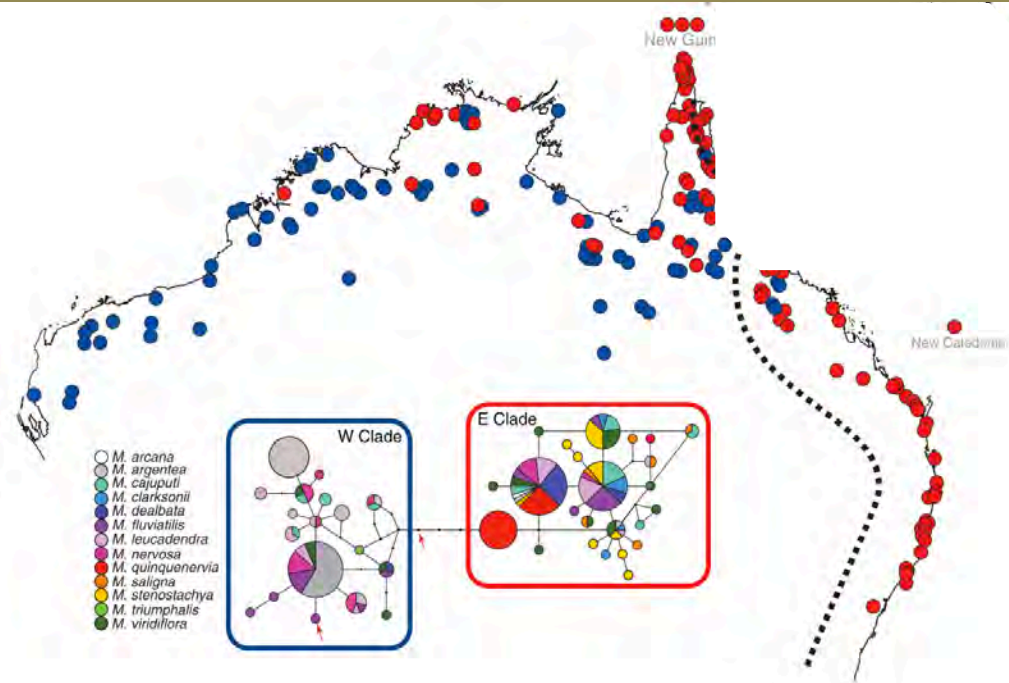
Systematics of Melaleuca - Broadleaved Paperbarks

CSIRO PUBLISHING

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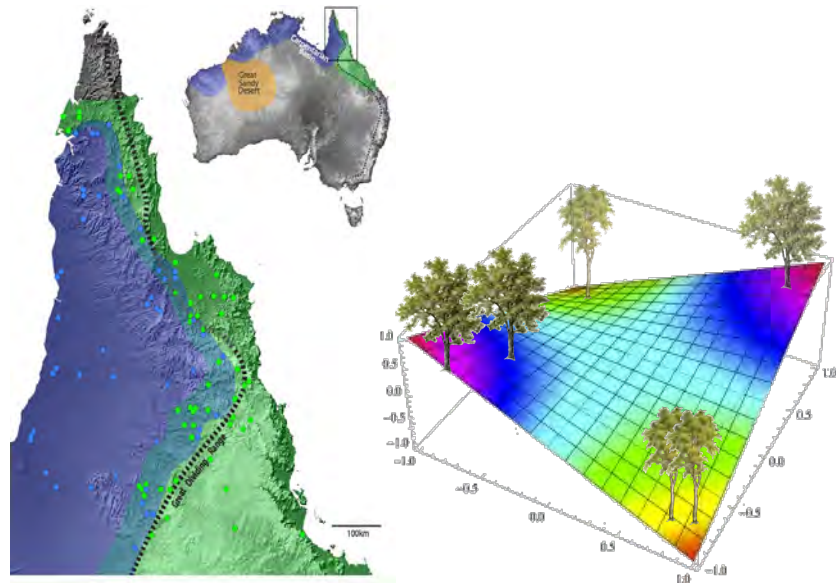


Journal of Biogeography (J. Biogeogr.) (2013) 40, 559–569

ORIGINAL
ARTICLE

Niche differentiation and spatial partitioning in the evolution of two Australian monsoon tropical tree species

Robert D. Edwards^{1*}, Michael D. Crisp² and Lyn G. Cook¹



- Niche values estimated via MaxEnt (Phillips *et al*, 2006)
- Values within and between species ranges are compared against randomized distributions (ENMtools (Glor & Warren, 2010))

Biogeography of the Australian Monsoon Flora

RESEARCH ARTICLE

Congruent biogeographical disjunctions at a continent-wide scale: Quantifying and clarifying the role of biogeographic barriers in the Australian tropics

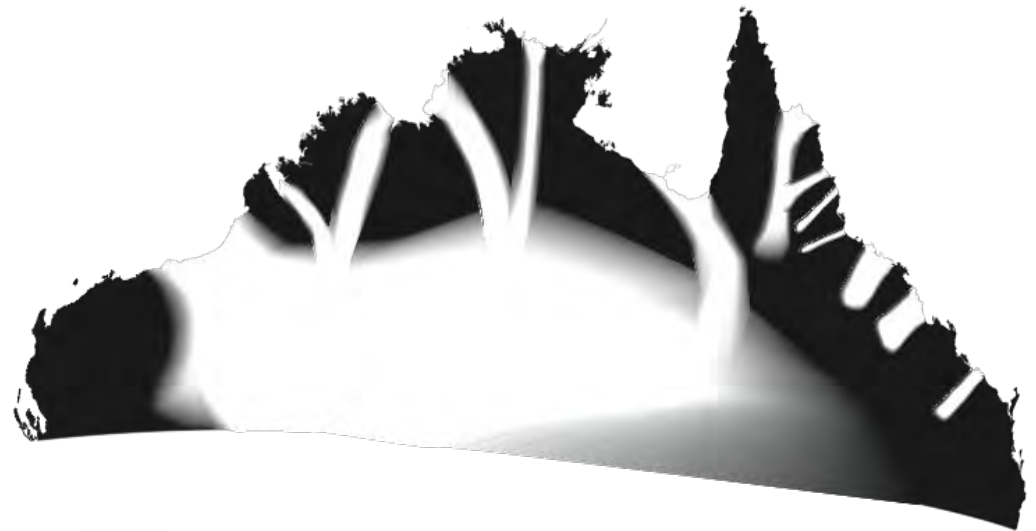
Robert D. Edwards^{1*}, Michael D. Crisp², Dianne H. Cook³, Lyn G. Cook¹

1 School of Biological Sciences, The University of Queensland, Brisbane, Queensland, Australia,

2 Research School of Biology, The Australian National University, Acton, Australian Capital Territory, Australia, **3** Department of Econometrics and Business Statistics, Monash University, Clayton, Victoria, Australia

* Current address: Department of Botany, National Museum of Natural History, MRC166, Smithsonian Institution, Washington, DC, United States of America

robertd4w@si.edu

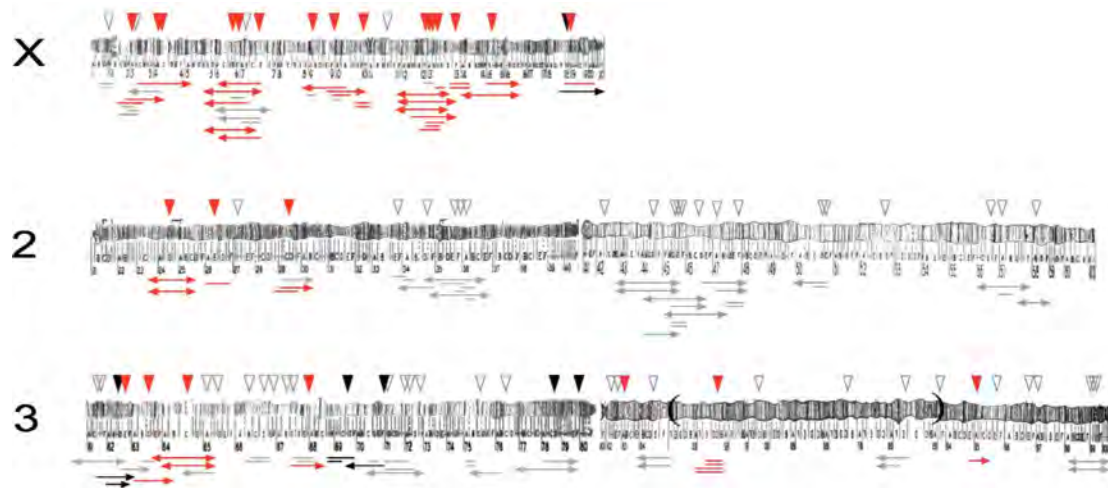




Species-isolation genes in *Drosophila*



UNIVERSITY of
ROCHESTER





Vicki Funk



Elisabeth Bui



Chase Mason



Powell Center



Marty Goldhaber



Ian Pearse



Travis Nauman



Jen Cartwright



Smithsonian
Institution



Brian Anacker



Joe Miller



Jim Thompson



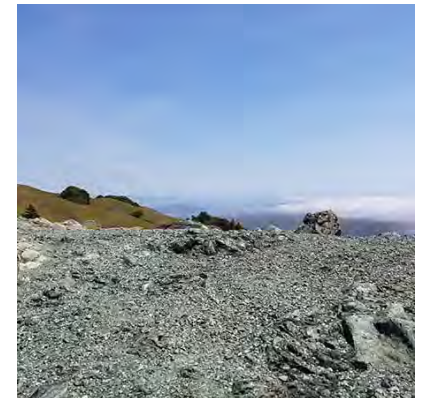
Pamela Soltis



Integrated Digitized Biocollections

Defining and Delimiting Extremes

Extreme environment: *a habitat where conditions are outside the optimal range for organisms to survive*





Defining and Delimiting Extremes

Extreme environment: *a habitat where conditions are outside the optimal range for organisms to survive*



Extreme habitats are often identified by a lack of organisms or lack of diversity of organisms



What for one organism is considered intolerable may not be so for another



Multiple factors may be “extreme” and contribute to the inhospitableness of an environment



As a result the threshold(s) for “extreme” are often idiosyncratic and determined on a by-species and/or by variable basis



Challenge:

Define extreme environments objectively

Test hypotheses on the response of biotic processes non-circularly
and across many species

- eg. trait presence, body size, species diversity



Defining and Delimiting Extremes

Challenge:

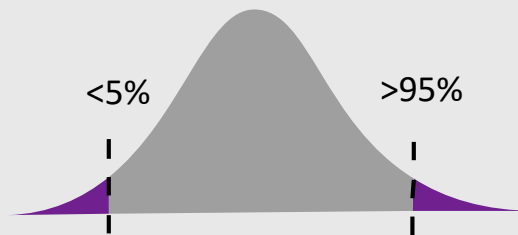
Define extreme environments objectively

Test hypotheses on the response of biotic processes non-circularly and across many species

– eg. trait presence, body size, species diversity

For terrestrial organisms the comfortable envelope for survival typically exists towards the middle of the range of environmental variables

ie. (by definition) the upper and/or lower tail of values for a variable are the extremes





Defining and Delimiting Extremes

Climate

Worldclim2 – 19 variables

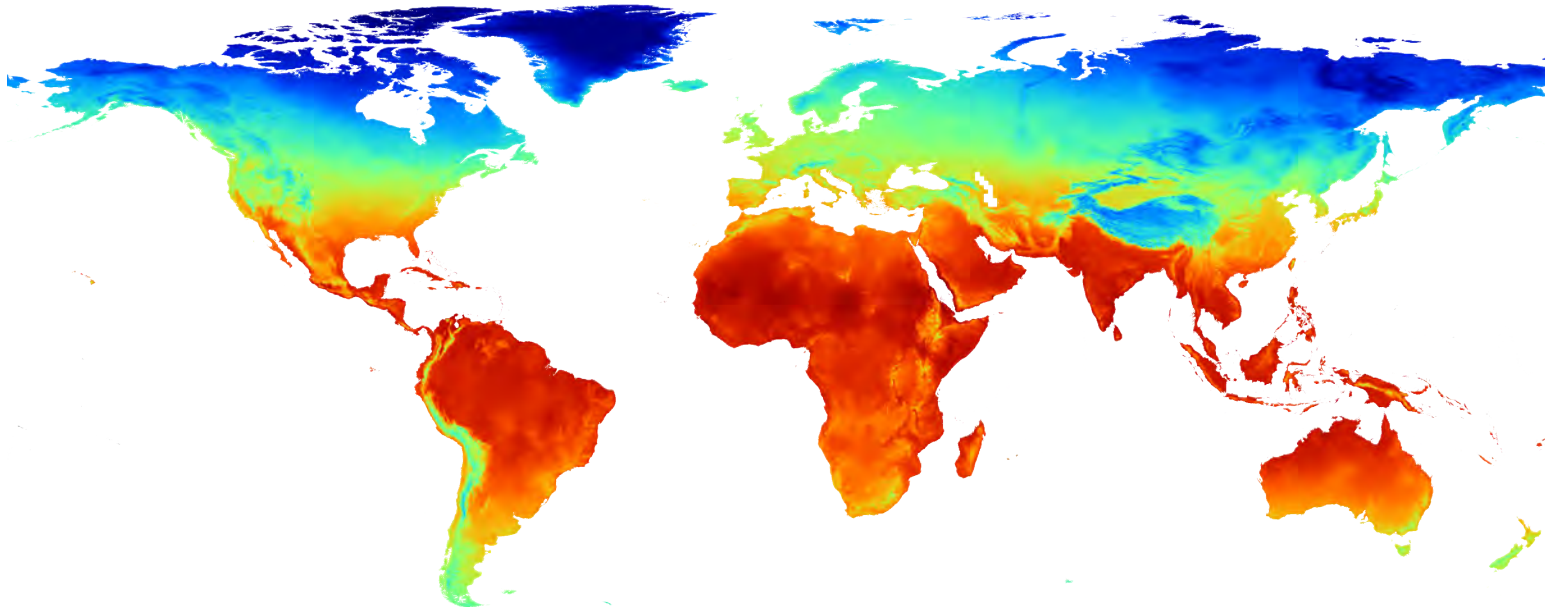
(Fick et al., 2017)

Soil and Elevation

SoilGrids 1km – 21 variables

(Hengl et al., 2017)

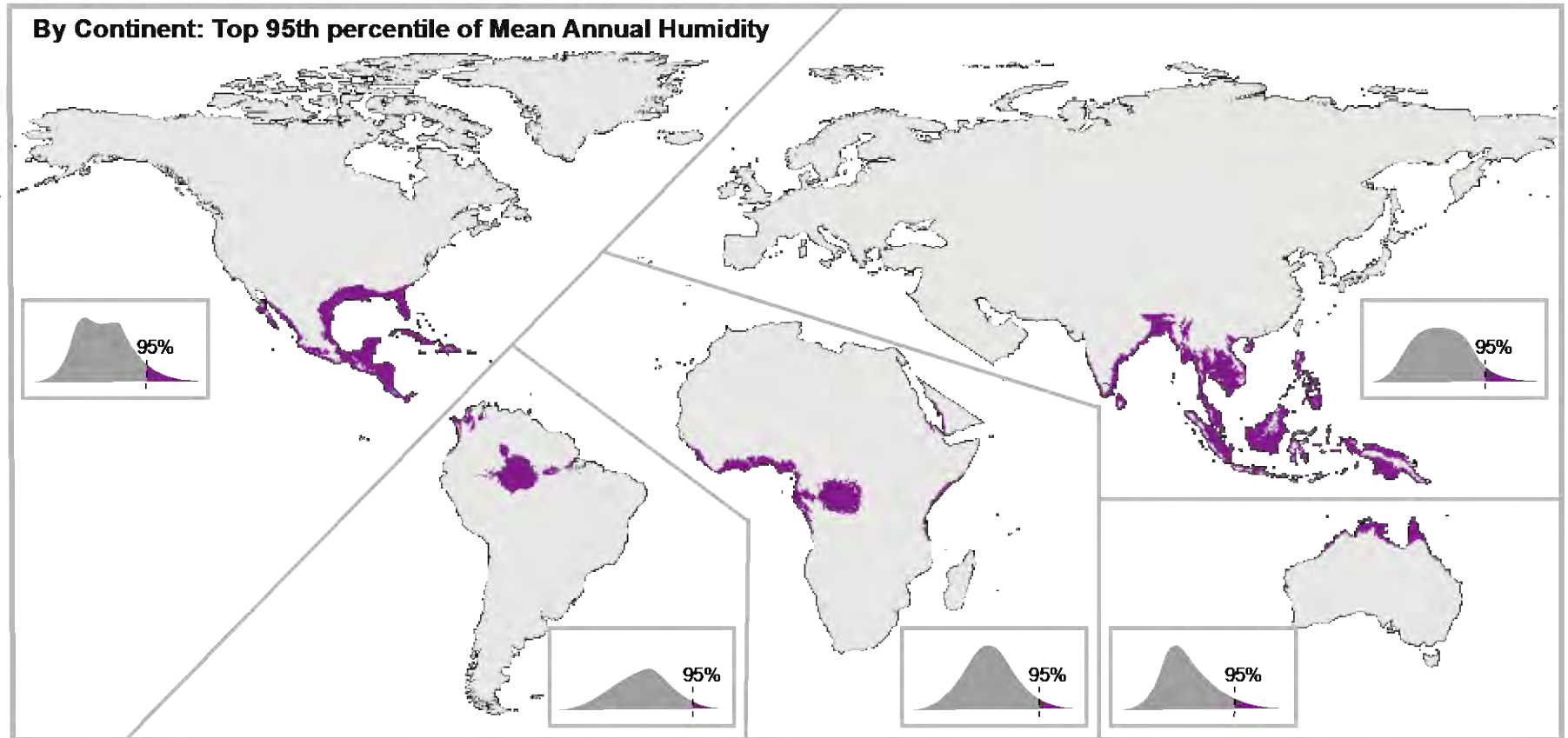
Initial 40 variables reduced to 24 relevant non-collinear set





Defining and Delimiting Extremes

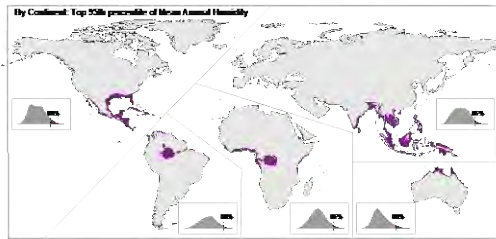
Individual variable



1km² pixel resolution



Defining and Delimiting Extremes



Stack of variables

Individual variable layer: Top 95th percentile of Topographic Wetness Index

Individual variable layer: Top 95th percentile of Annual Precipitation

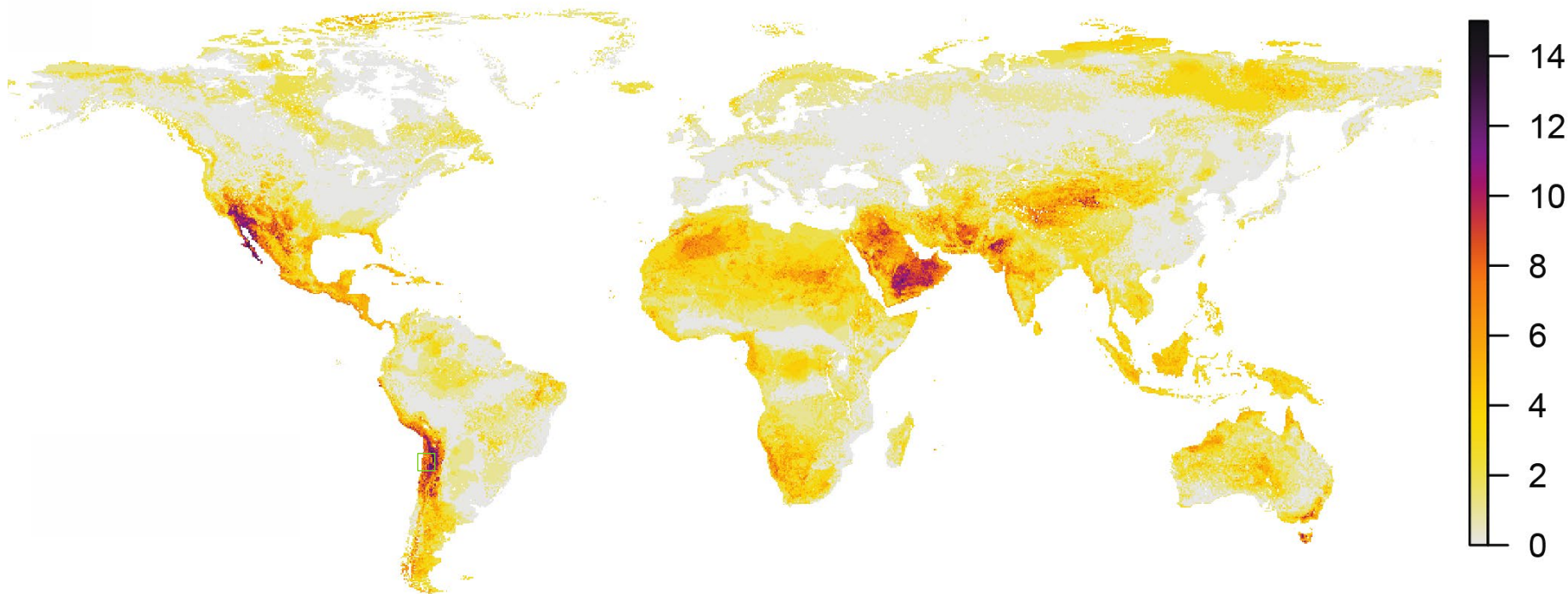
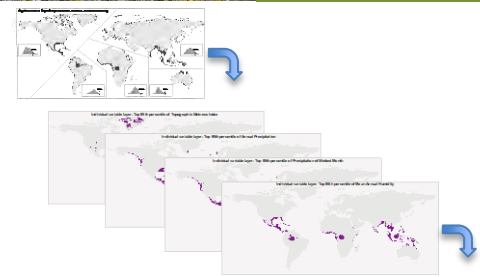
Individual variable layer: Top 95th percentile of Precipitation of Wettest Month

Individual variable layer: Top 95th percentile of Mean Annual Humidity



Daisy Diversity Response to Extremes

- Defining and Delimiting Extremes



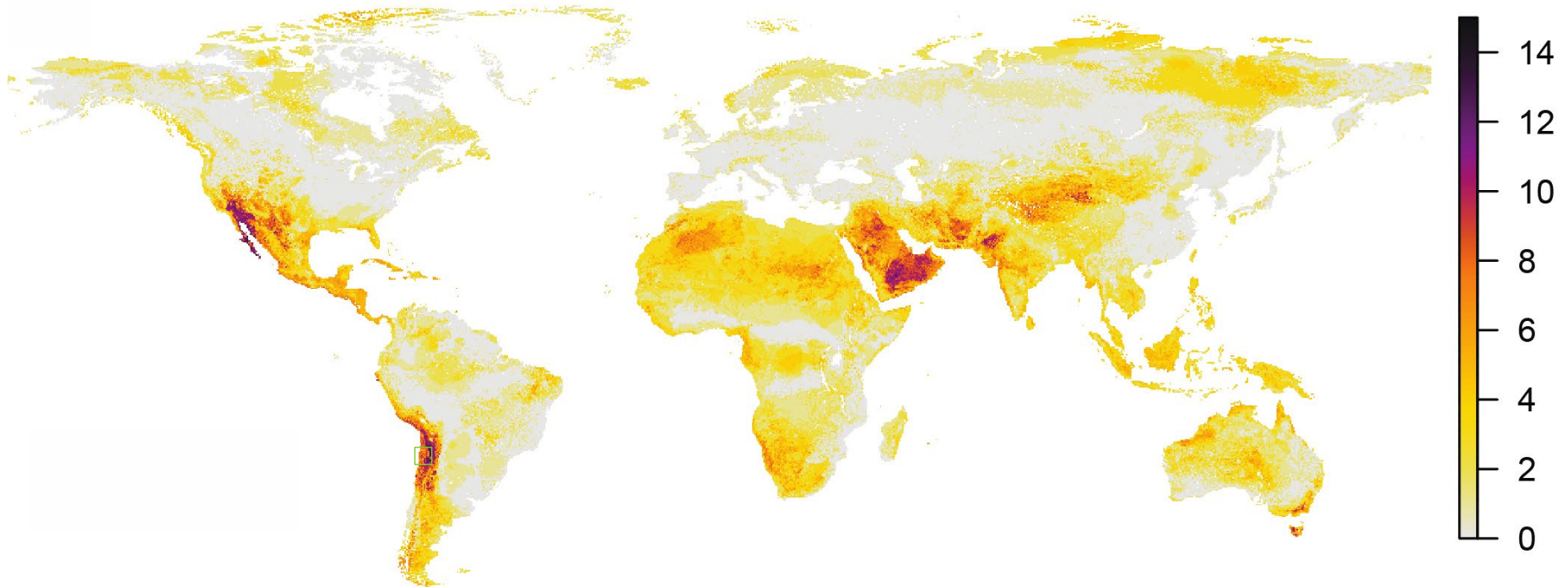
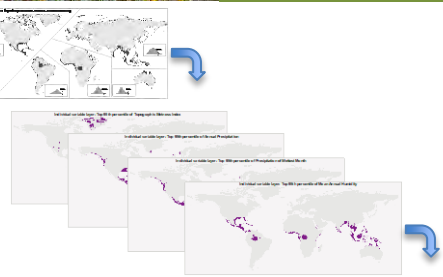
Number of Overlapping Extremes



Daisy Diversity Response to Extremes

- Defining and Delimiting Extremes

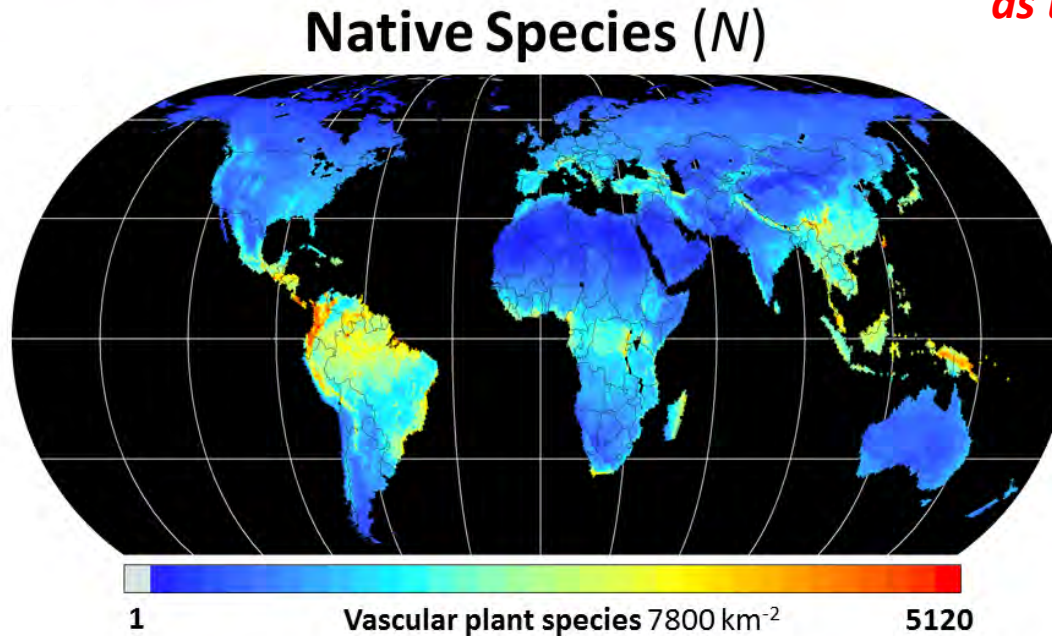
Hypothesis:
*Species diversity will decrease
as the number of overlapping
extremes increases*



Number of Overlapping Extremes



Hypothesis:
*Species diversity will decrease
as the number of overlapping
extremes increases*



Ellis et al., 2012; after Kreft and Jetz, 2007.

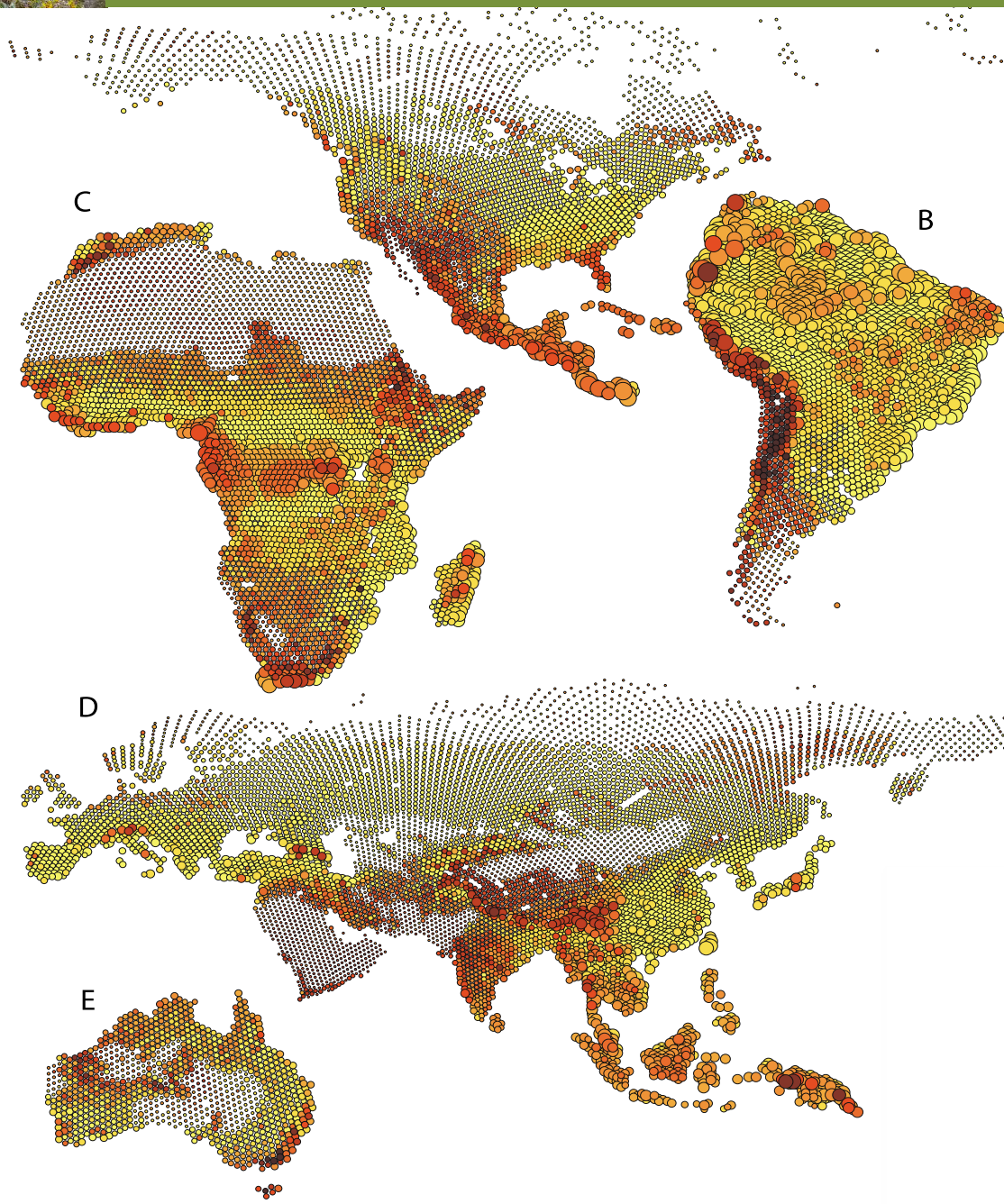
- 16,461 equal area polygons each with a value for flowering plant richness
- Value for the number of extremes also recorded for the centroid of each polygon



Flowering Plant Diversity Response to Extremes

- Global

Hypothesis:
*Species diversity will decrease
as the number of overlapping
extremes increases*



Richness

- 0 - 512
- 513 - 1024
- 1025 - 1536
- 1537 - 2048
- 2049 - 2560
- 2561 - 3072
- 3073 - 3584
- 3585 - 4096
- 4097 - 4608
- 4609 - 5121

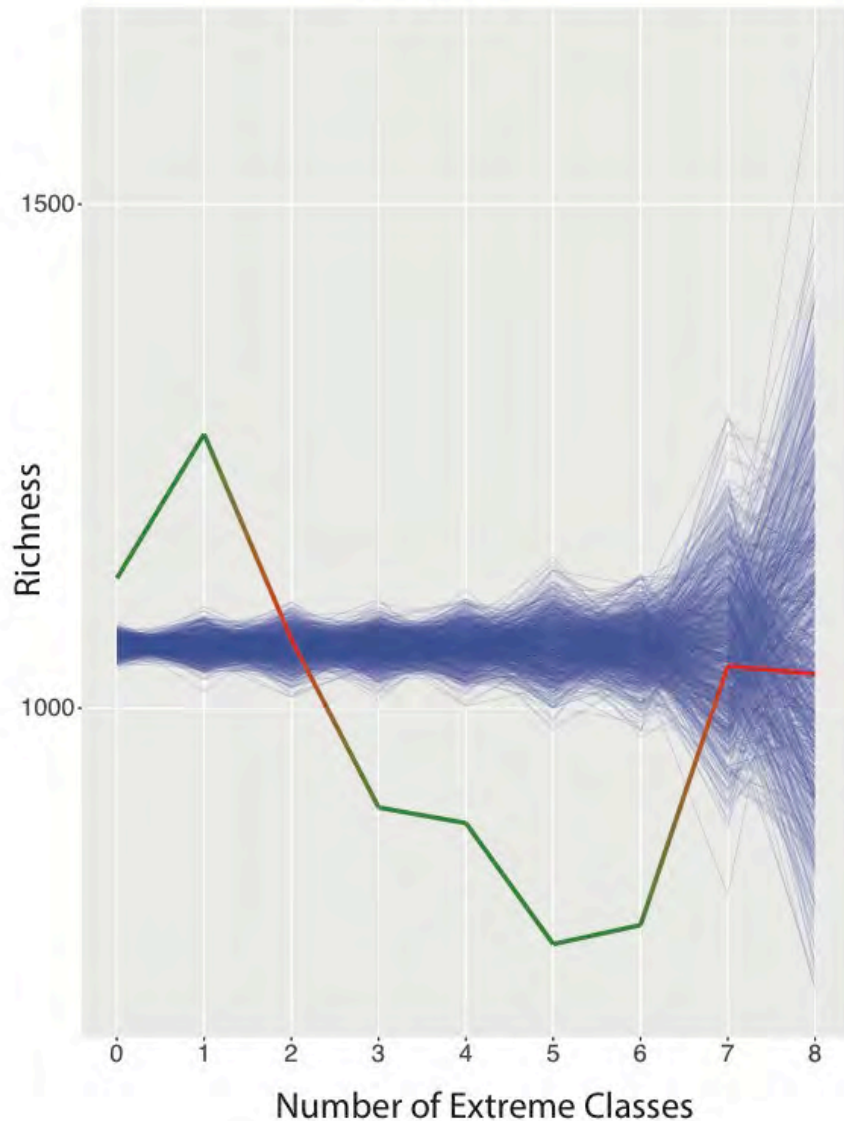
Extremeness

- 0
- 1
- 2
- 3
- 4
- 5
- 6
- 7
- 8



Hypothesis:
*Species diversity will decrease
as the number of overlapping
extremes increases*

Global



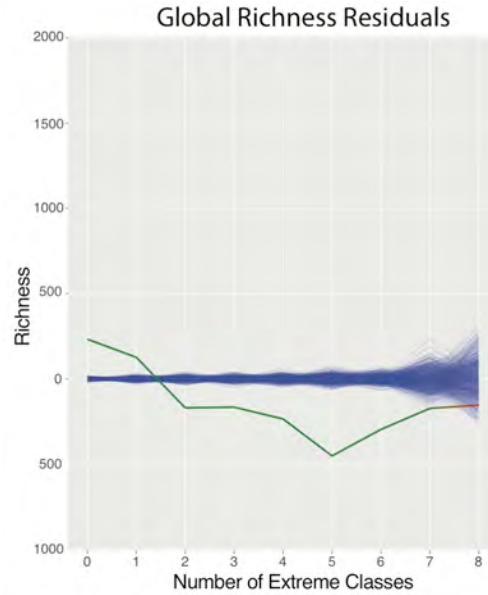
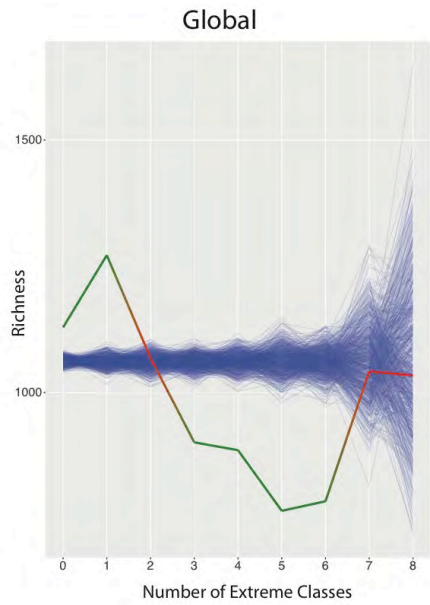
Calculate the average species richness for each number of overlapping classes

Null distribution generated from 999 Randomizations of richness values for each grid cell while holding number of overlapping classes constant



Global Flowering Plant Diversity Response to Extremes

Latitudinal Species Gradient?

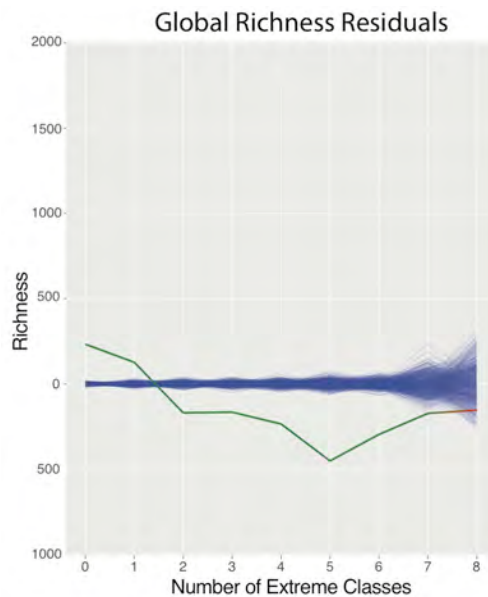
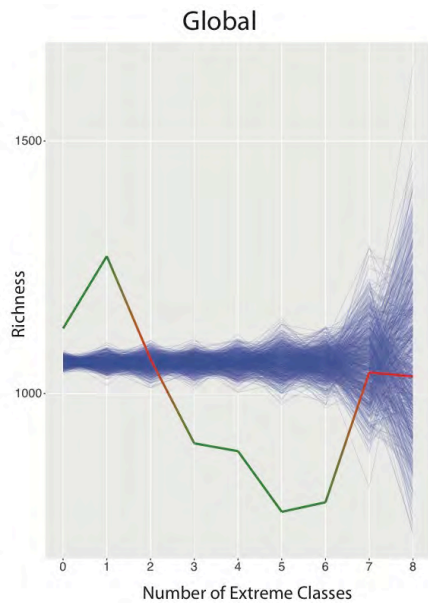


Regression of number
of extreme classes
against latitude:
 $R^2 < 0.145$

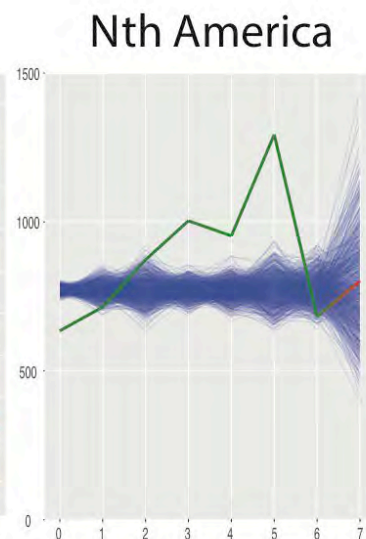
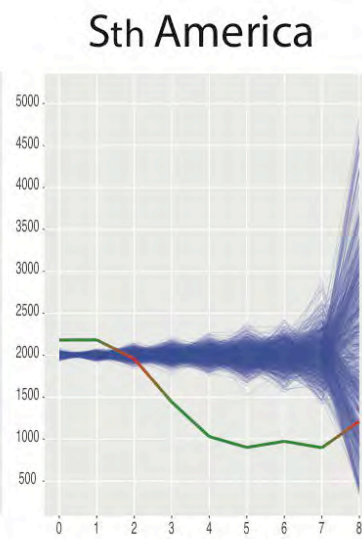
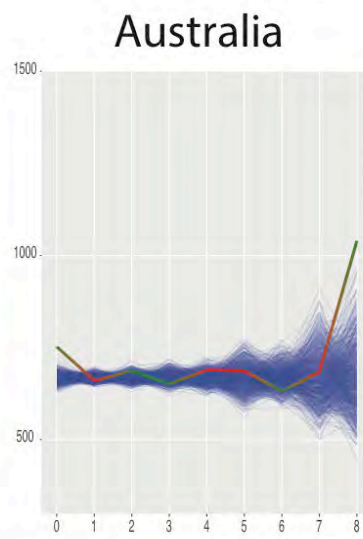
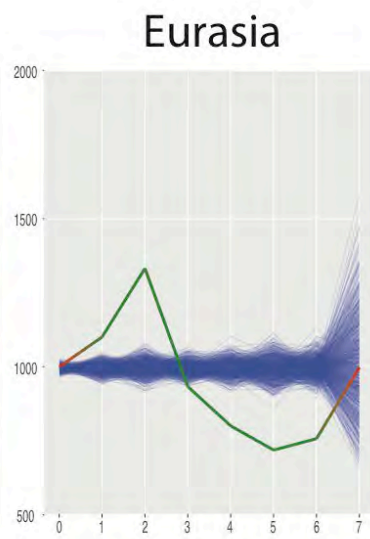
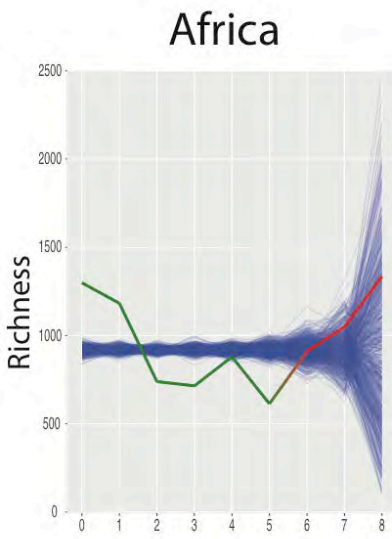


Global Flowering Plant Diversity Response to Extremes

Latitudinal Species Gradient?



Regression of number
of extreme classes
against latitude:
 $R^2 < 0.145$

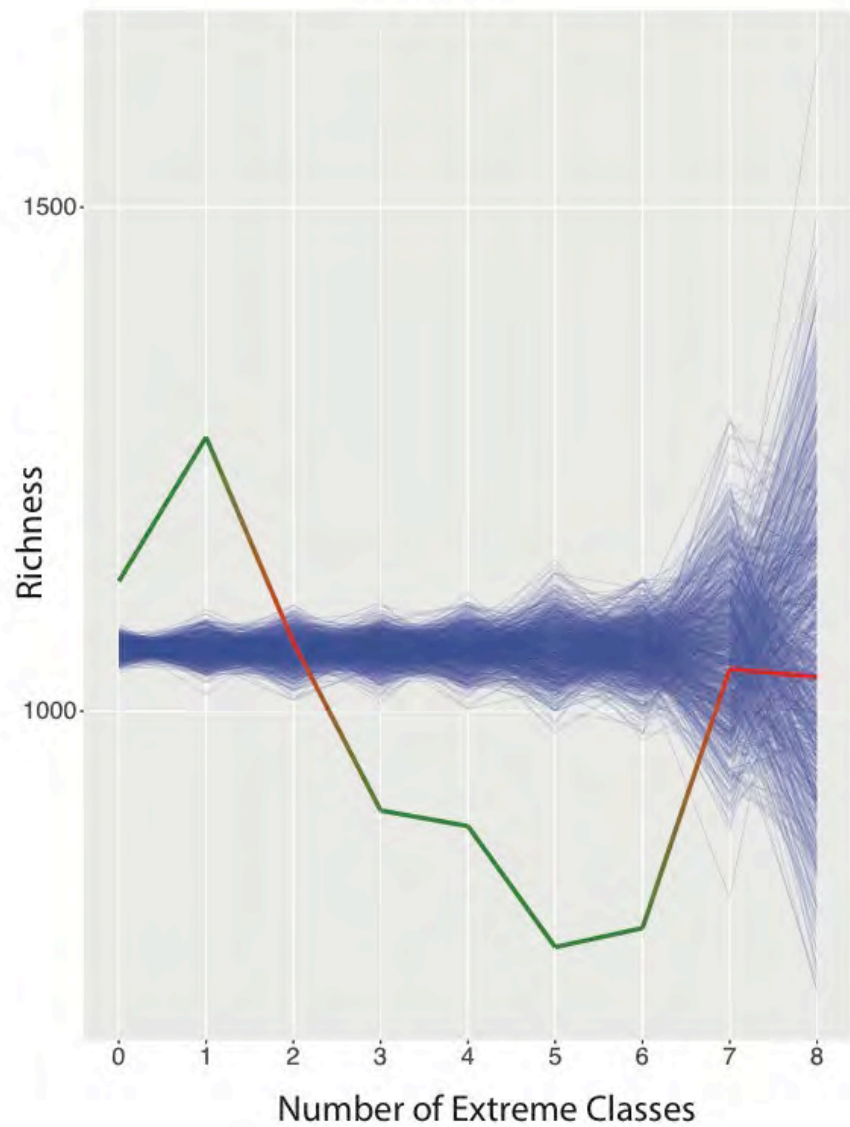


Number of extremes

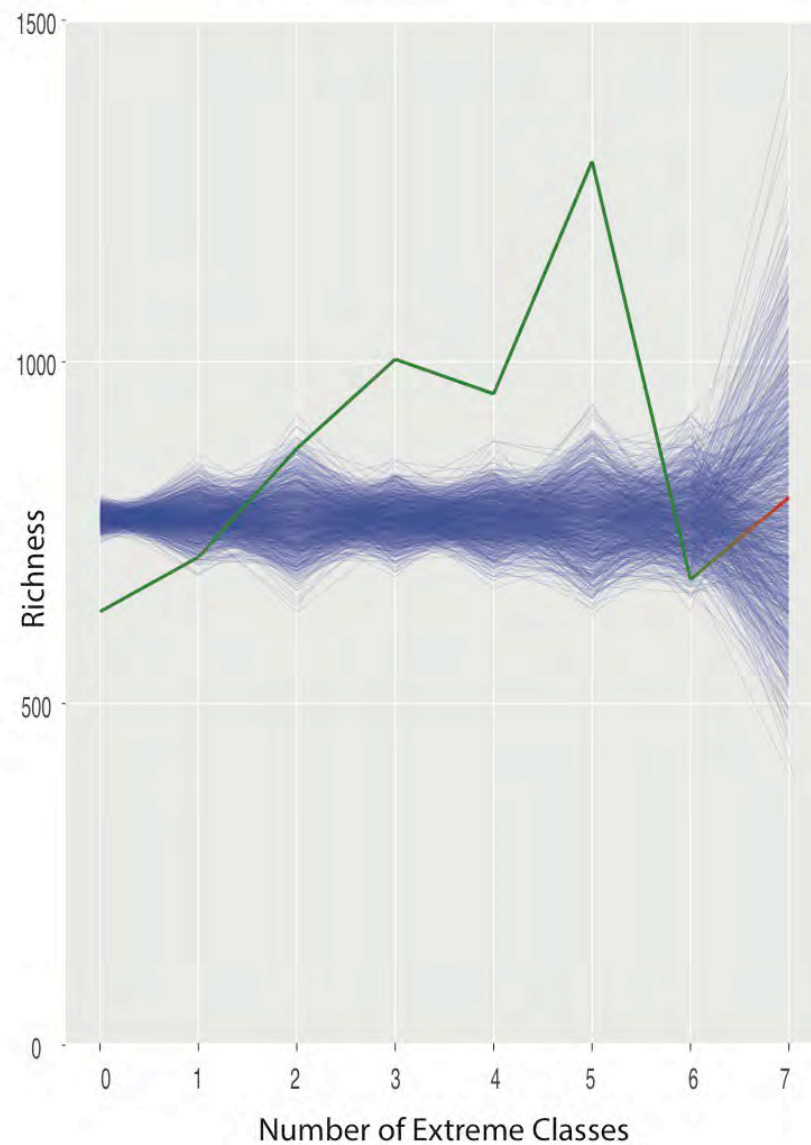


Flowering Plant Diversity Response to Extremes

Global



Nth America
All Flowering Plants





North American Compositae (Asteraceae)

Daisies, sunflowers, thistles, tarweeds etc

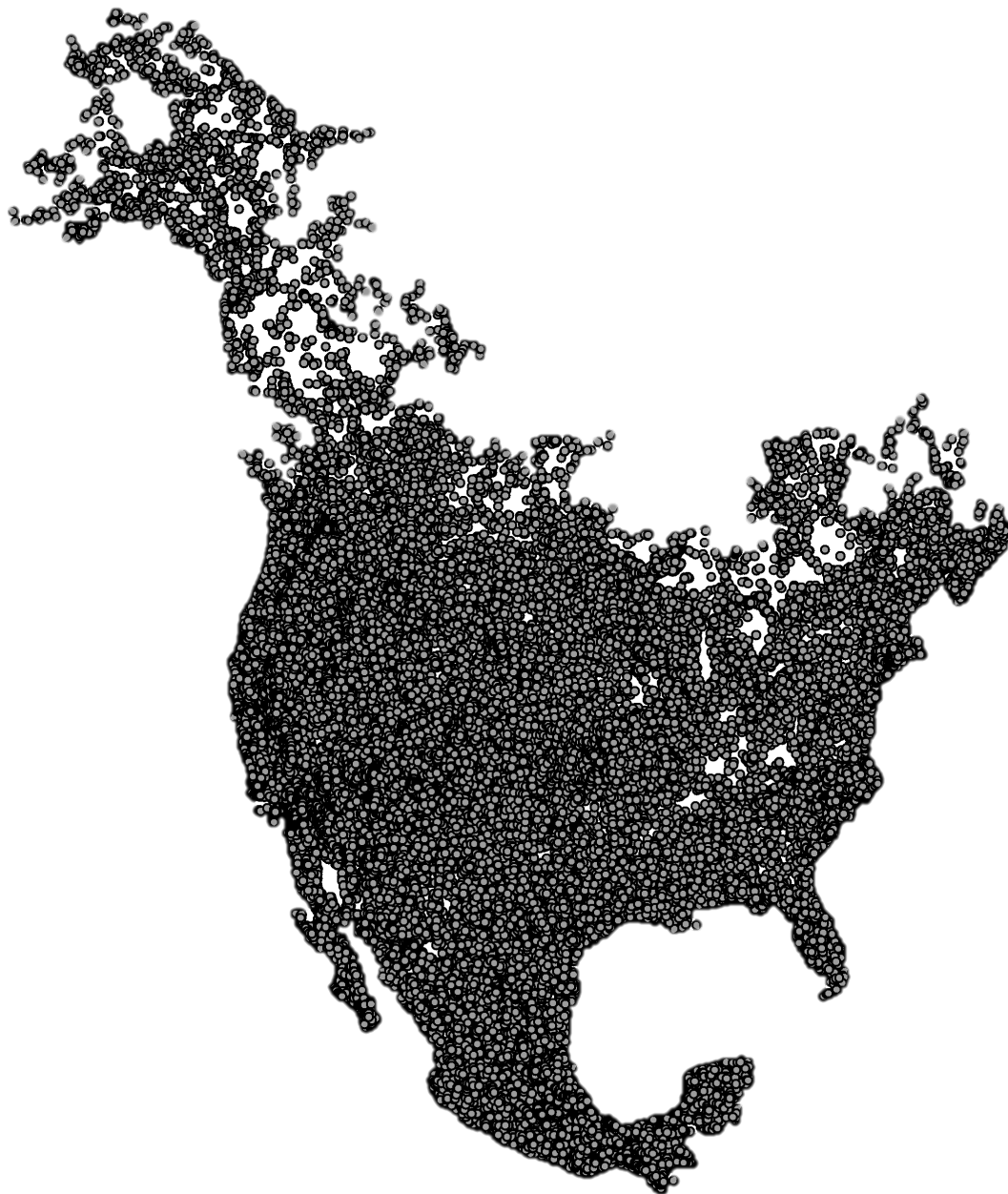
Massively diverse (~1 in 10 flowering plants)

Known for occupying a wide range of challenging environments



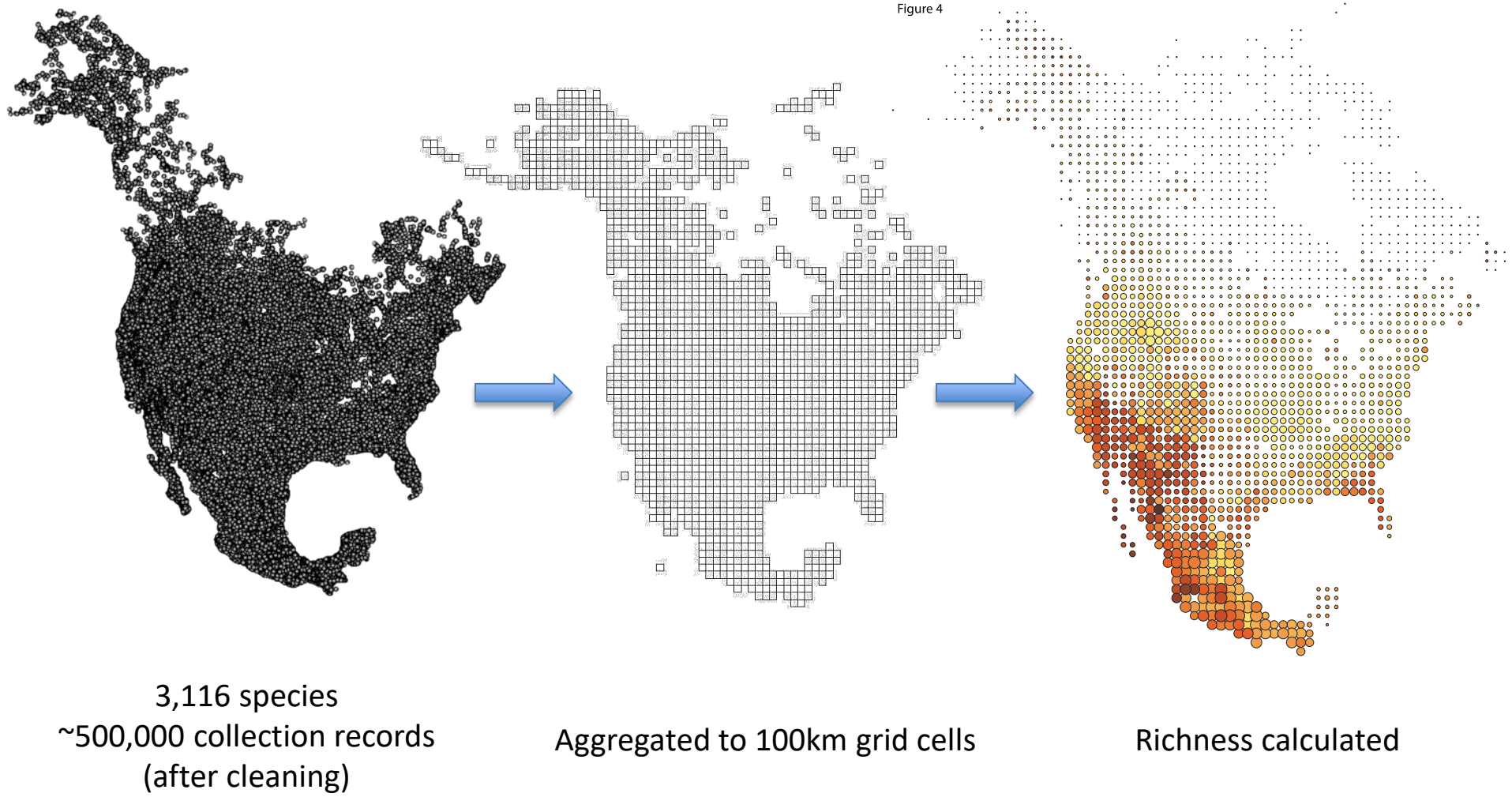


Daisy Diversity Response to Extremes





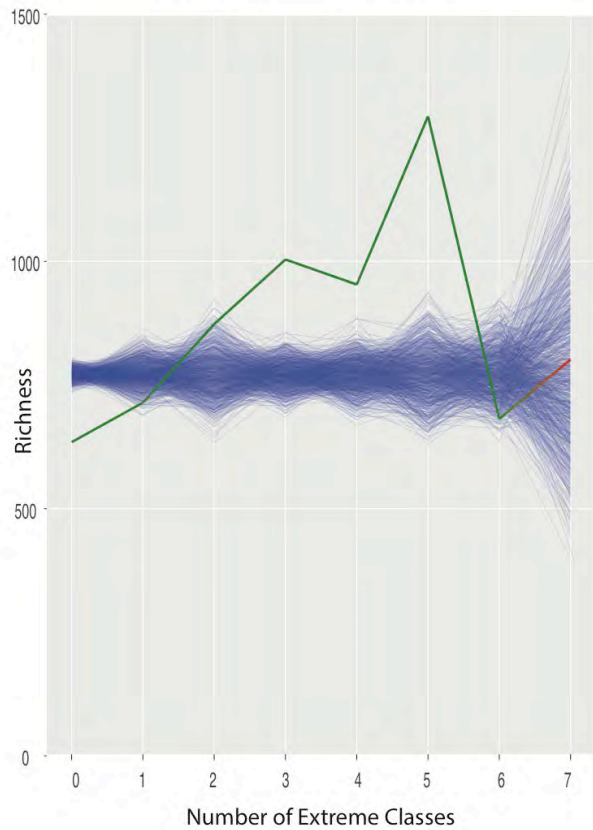
Daisy Diversity Response to Extremes





Daisy Diversity Response to Extremes

Nth America
All Flowering Plants

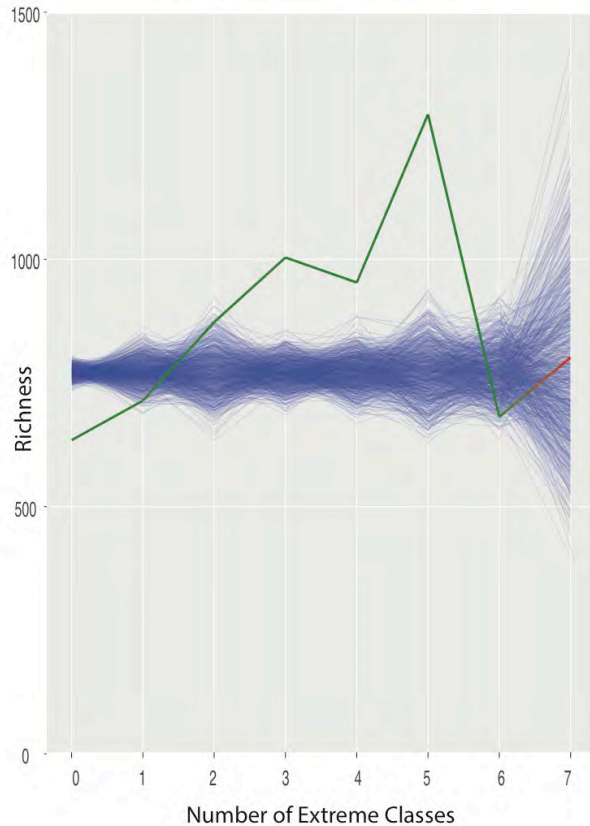


In Nth America species richness increases with extremeness...

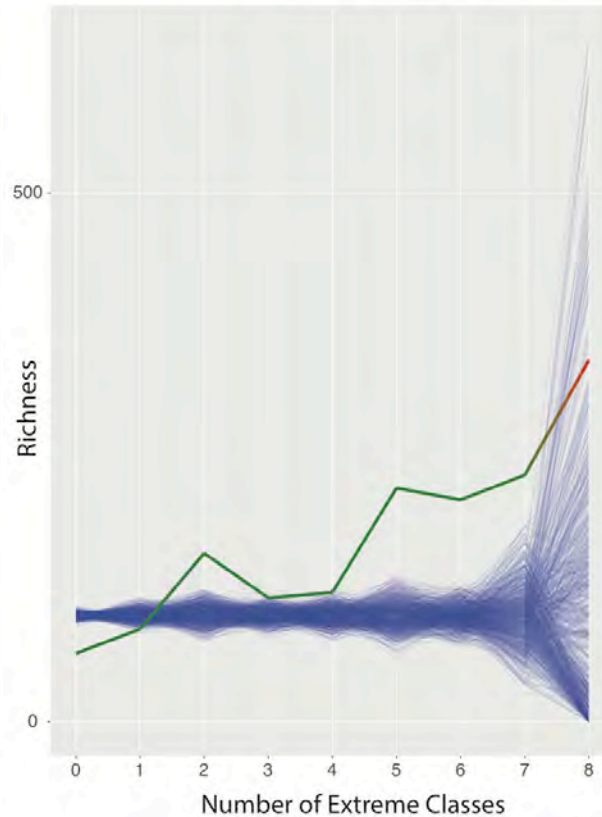


Daisy Diversity Response to Extremes

Nth America
All Flowering Plants



Nth American Compositae



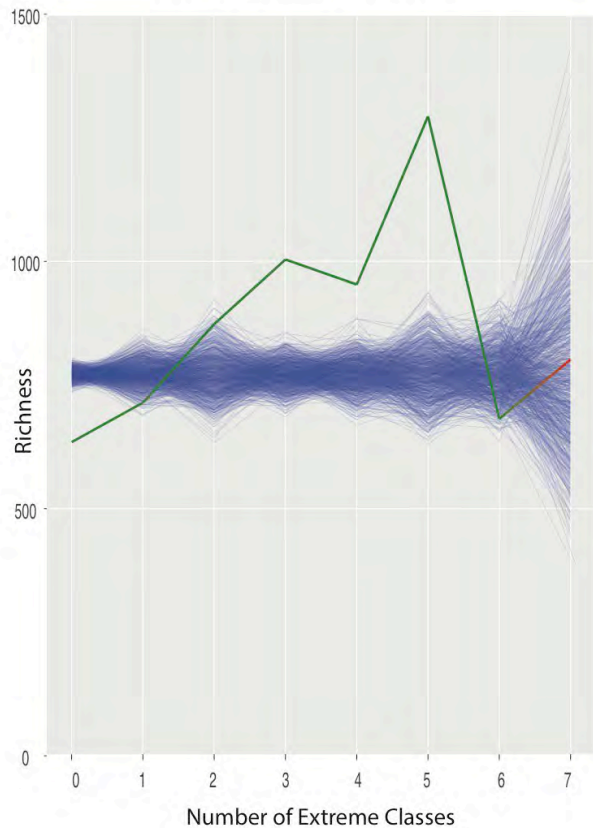
In Nth America species richness increases with extremeness...

...this trend is especially prominent in the daisy family...

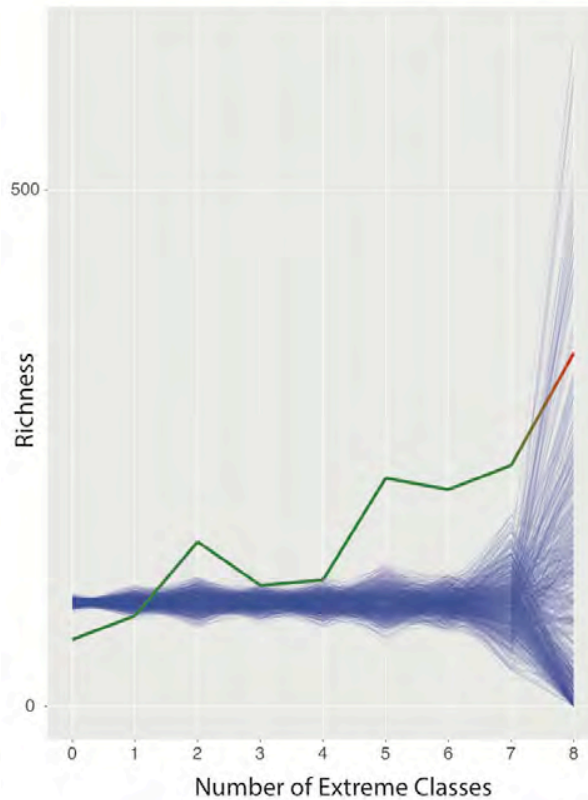


Daisy Diversity Response to Extremes

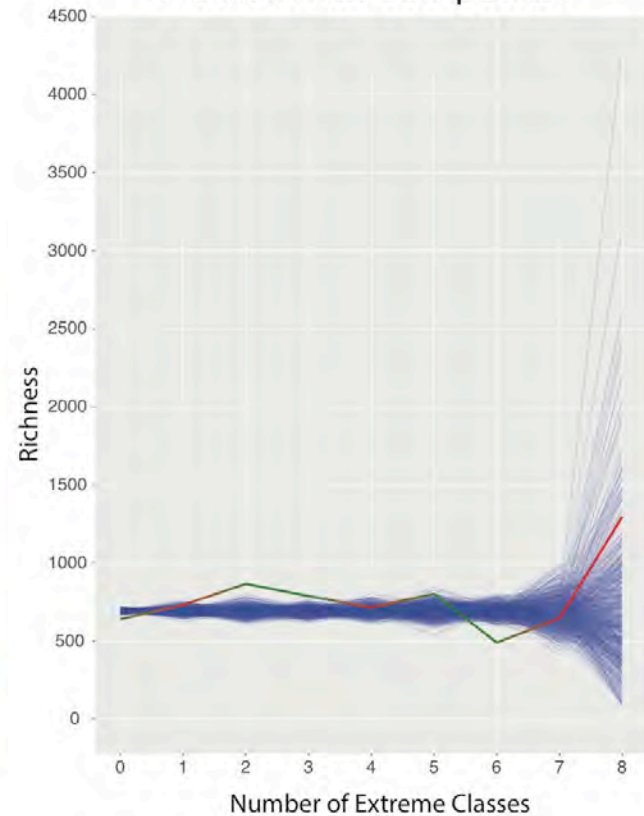
Nth America
All Flowering Plants



Nth American Compositae



Nth American Flowering
Plants Minus Compositae



In Nth America species richness increases with extremeness...

...this trend is especially prominent in the daisy family...

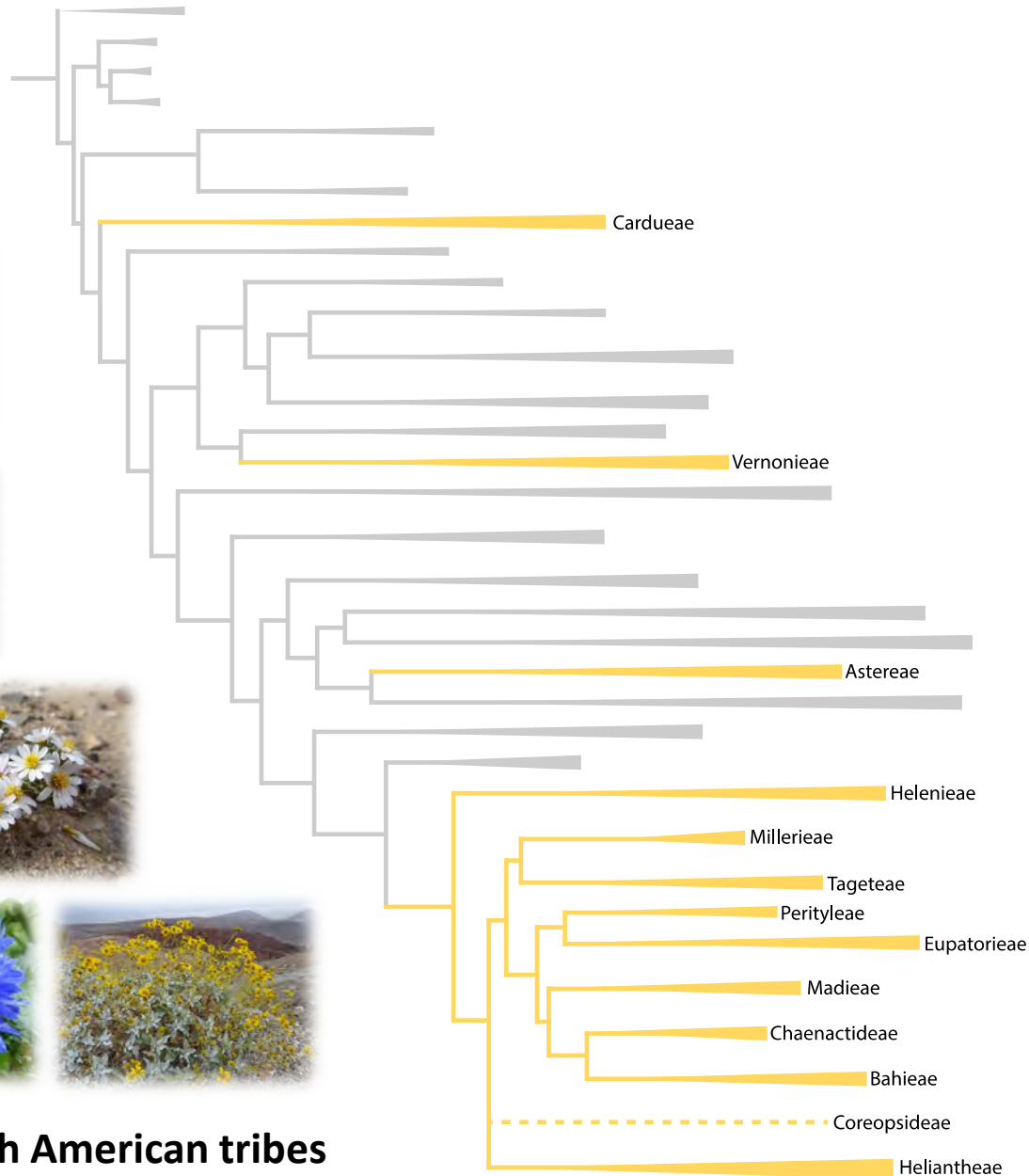
...to the point that they account for almost all of it!



Drivers of Diversification



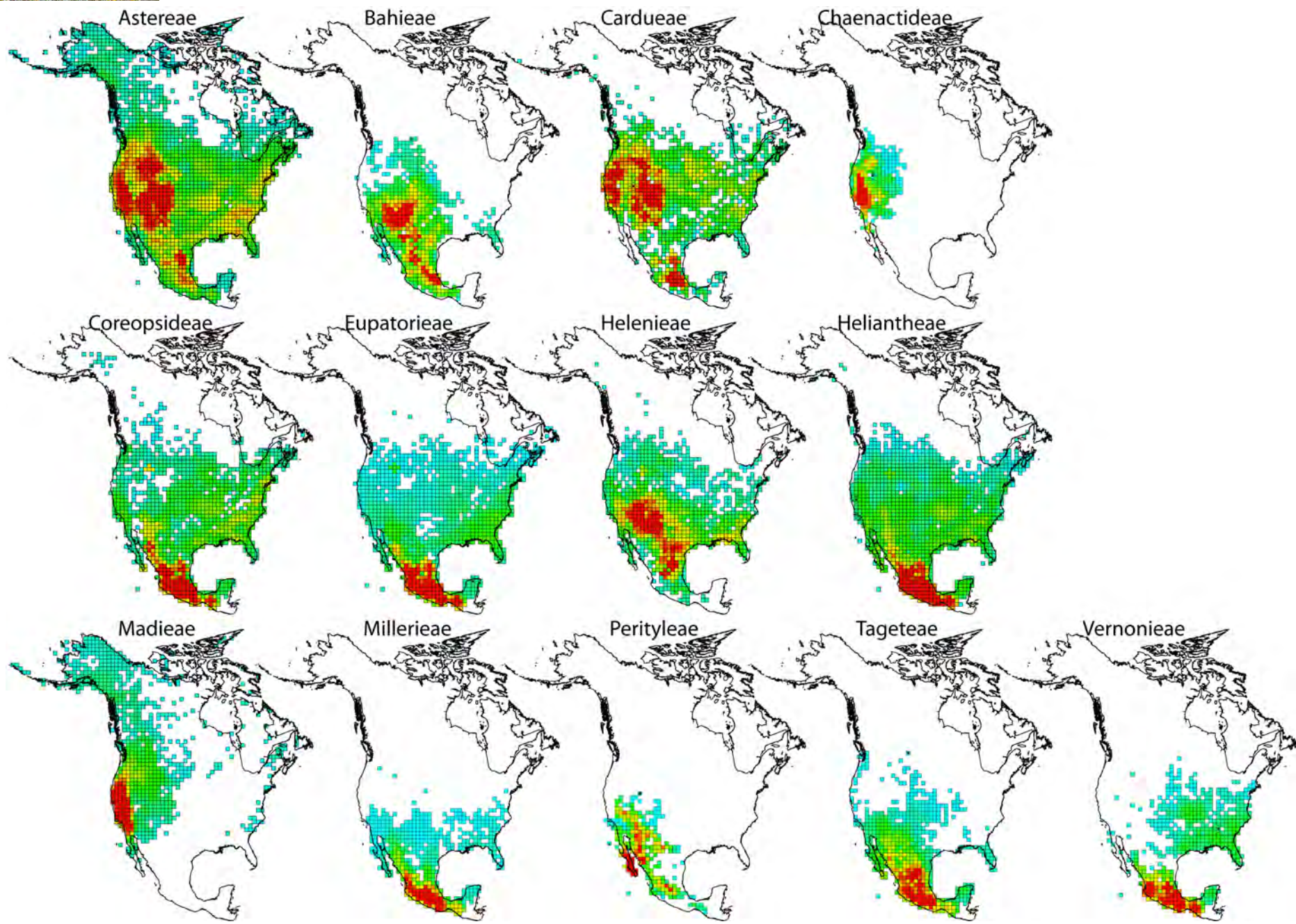
13 majority endemic Nth American tribes
4 ~independent radiations



Modified from Mandel et al., 2019



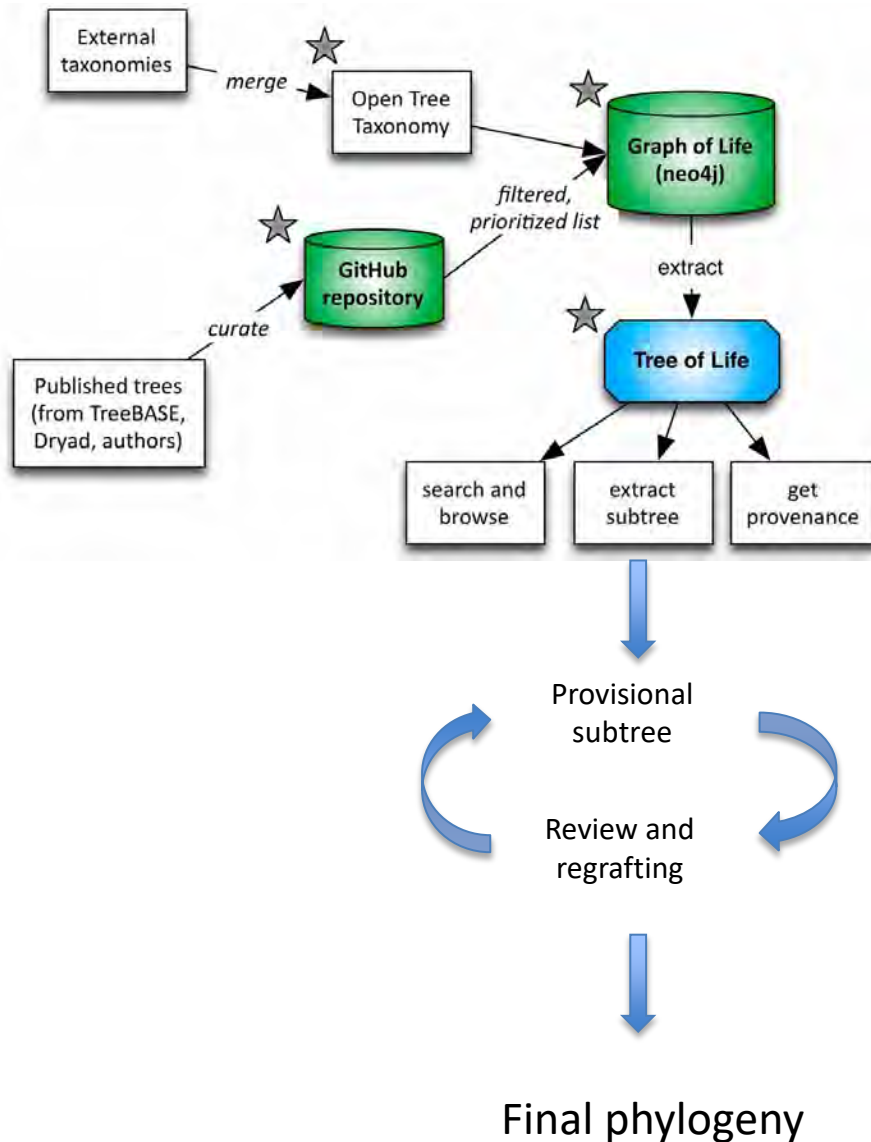
Daisy Diversity Response to Extremes





Drivers of Diversification

Generating a species-level meta-phylogeny



Open Tree of Life
+
GenBank
+
Expert assessment and regrafting

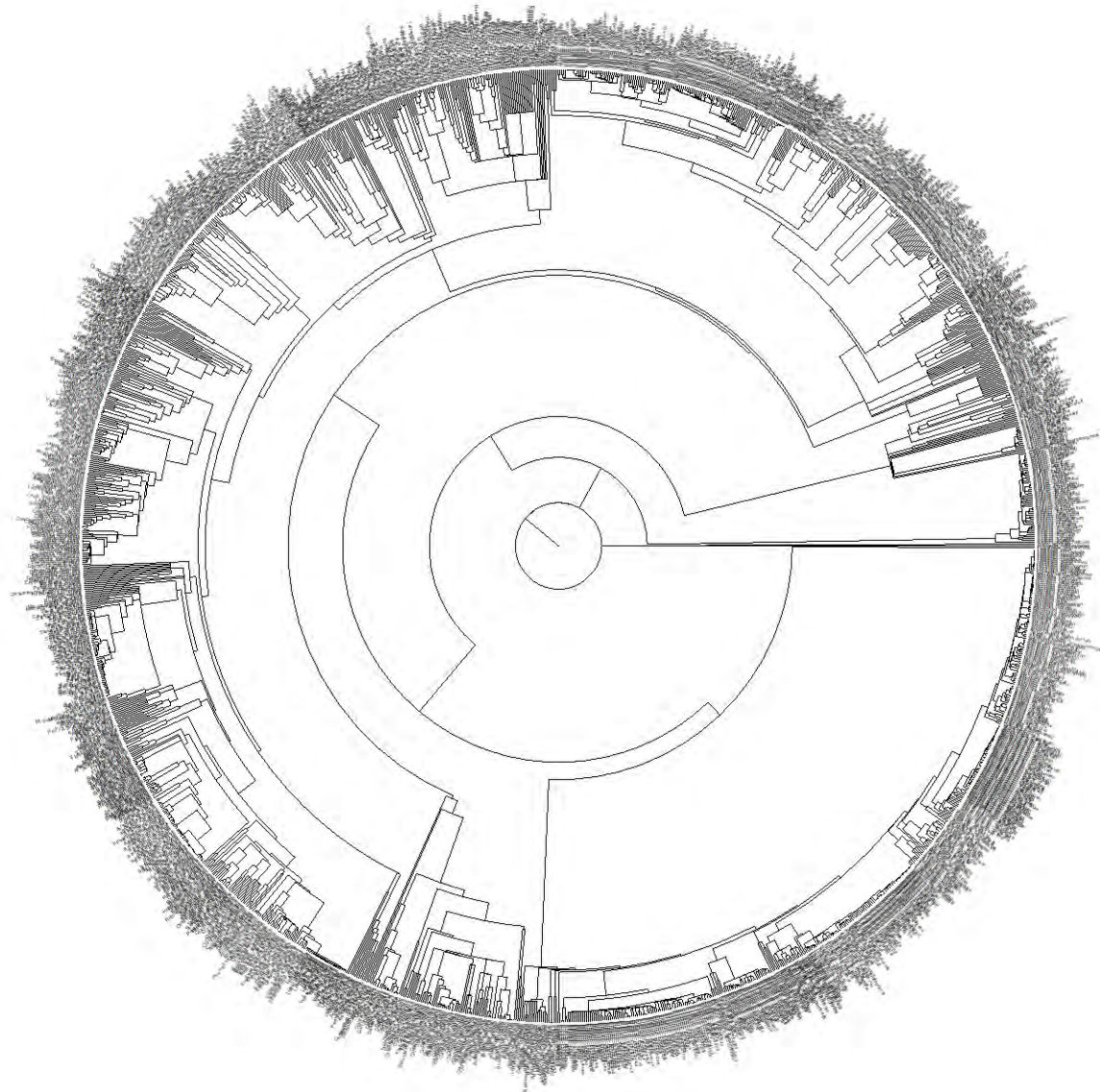
- Polytomies
- Misclassifications
- Missing taxa



Stephen Smith

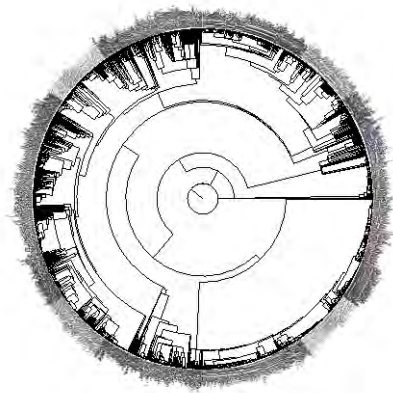


Drivers of Diversification

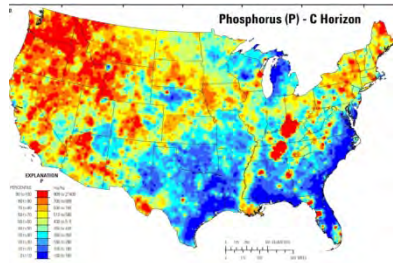


1642 taxa

Drivers of Diversification



Which environmental variables are associated with increased/decreased diversification?



Climate – AdaptWest 1km (ClimateNA v5.10)¹

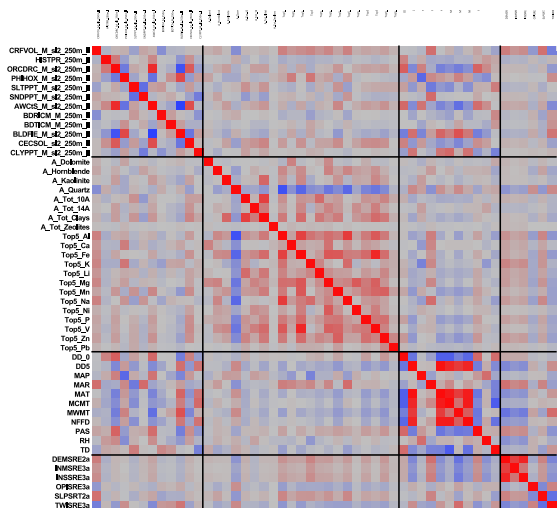
Soil Profile/Texture – SoilGrids 250m v0.5²

Topography – DEM-derived, via SoilGrids²

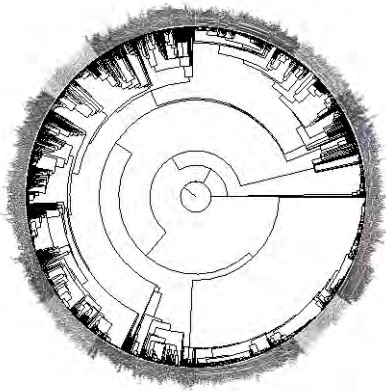
Soil Geochemical/Mineralogical Data

– USGS (OFR 2014-1082)³ [interpolated from 4800 sites]

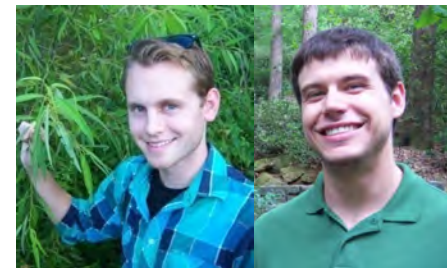
- 182 variables reduced to 50 non-collinear



“Variable reduction is a truly dark art” – Voldemort (probably)



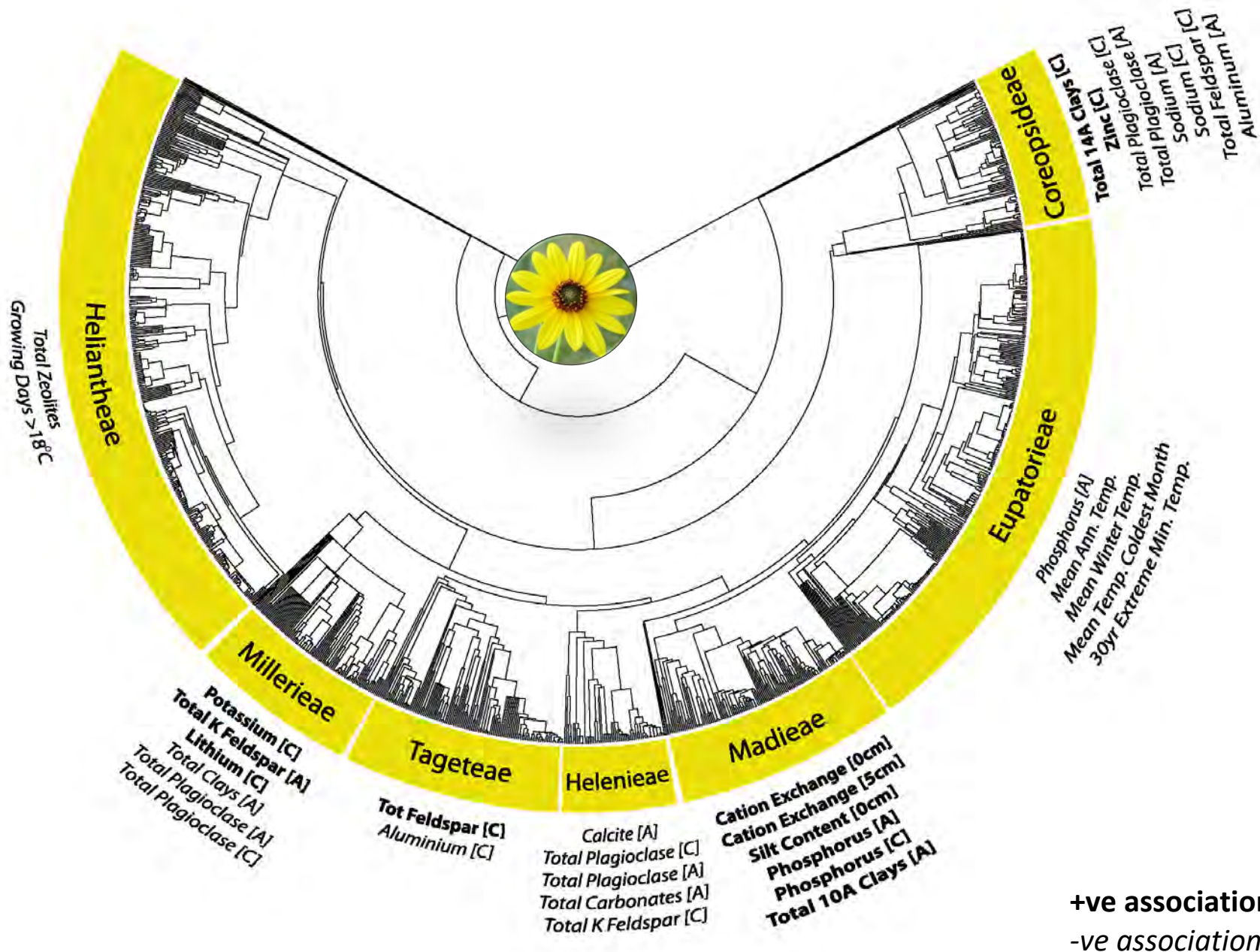
- Tip-rate correlation tests using the Diversification Rate metric (Jetz et al., 2012)
 - Ordinary Least Squares regression (to meet assumptions and proven better performance over PGLS - Revell, 2010)
 - First regressed diversification rate against single environmental variables
-
- Then multiple regression with all pairwise combinations of environmental variables that were not autocorrelated ($r > 0.75$), including additive ($A + B$) and multiplicative ($A * B$).
 - To control for false positives, we replaced our environmental data with randomly generated data (x1000) and repeated the analysis process for each simulated dataset.
 - R^2 values were pooled across runs, and the 95th percentile was used as a cutoff value for interpreting the reliability of results with real data.



Chase Mason & Eric Goolsby



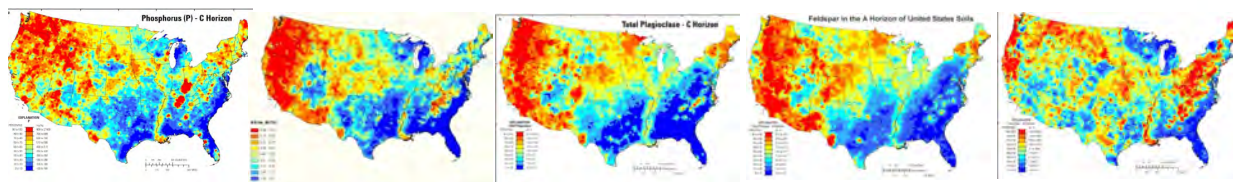
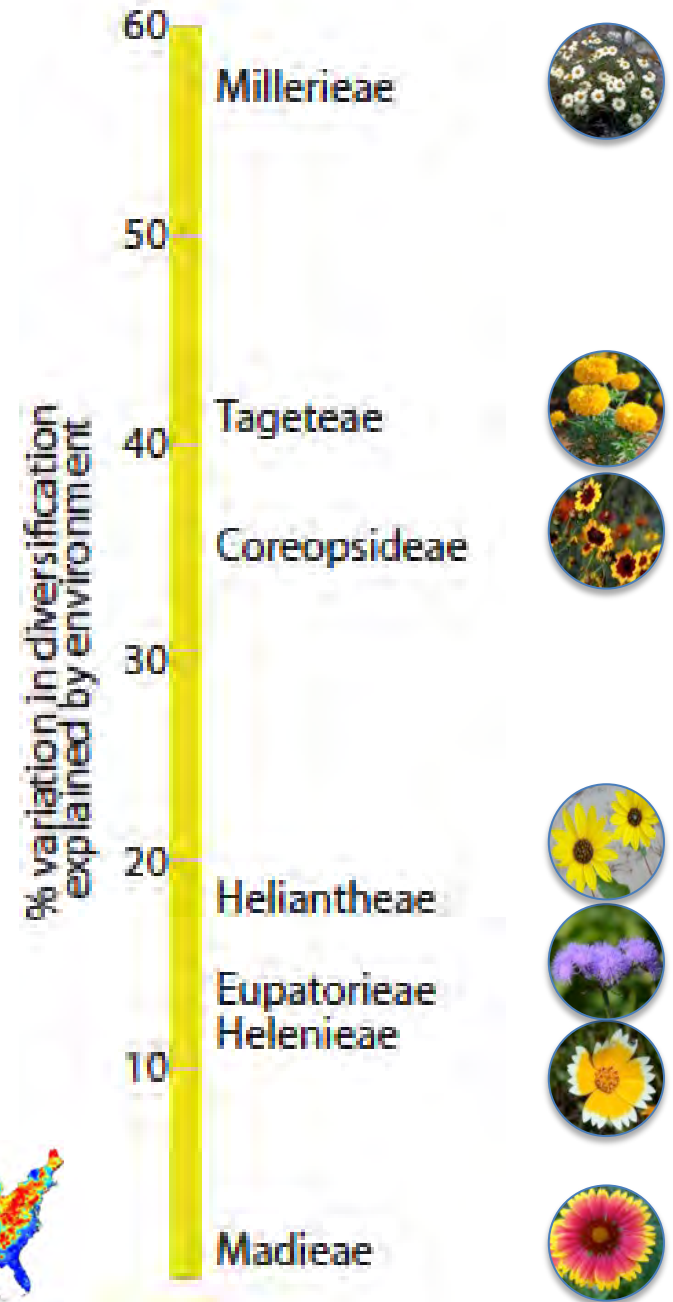
Drivers of Diversification



What factors drive diversification across North American Asteraceae?

- Abiotic environmental factors can explain >50% of variation in diversification rate, and up to 30% above randomly simulated data.
- This varies substantially among tribes.
- As a variable class, geochemical and mineralogical factors are consistently more correlated with diversification rate than climate, topography, or soil profile/texture:

*Phosphorus, Potassium, Aluminum
Plagioclases, Feldspars, Clays*



Global Vascular Plant Master Taxonomy

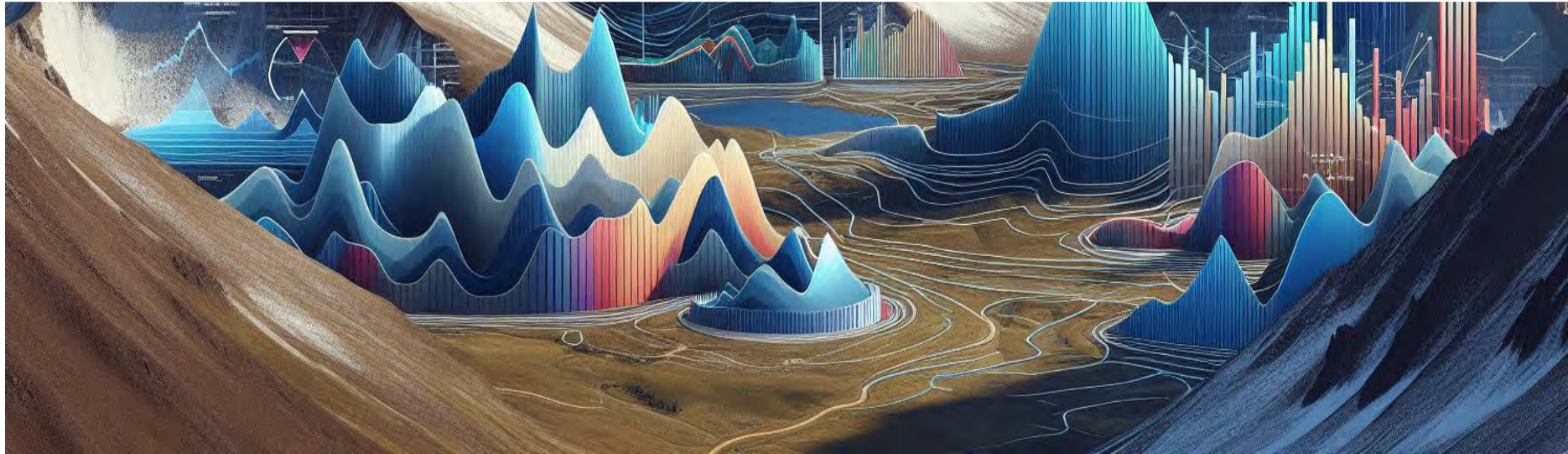
Enkleia paniculata Hallier f. | Enkleia | paniculata | NA | NA | TRUE | Hallier f. | Enkleia paniculata | NA | synonym | Enkleia paniculata (Merr.) Hallier f. | species | GBIF | Author Variant | Enkleia par
Enkleia riouwensis Hallier f. | Enkleia | riouwensis | NA | NA | TRUE | Hallier f. | Enkleia riouwensis | NA | synonym | Enkleia malaccensis Griff. | species | LCVF | Enkleia malaccensis | Enkleia | malacc
Enkleia siamensis (Kurz) Nevling | Enkleia | siamensis | NA | NA | TRUE | (Kurz) Nevling | Enkleia siamensis | 1961 | synonym | Enkleia malaccensis Griff. | species | LCVF | Enkleia malaccensis | Enkleia | m
Enkleia siamensis subsp. andamanica (Hutch. ex C. E. Parkinson) Nevling | Enkleia | siamensis | subsp. | andamanica | TRUE | (Hutch. ex C. E. Parkinson) Nevling | Enkleia siamensis subsp. andam
Enkleia siamensis subsp. andamanica (Hutch. ex C. E. Parkinson) Nevling | Enkleia | siamensis | subsp. | andamanica | TRUE | (Hutch. ex C. E. Parkinson) Nevling | Enkleia siamensis subsp. andaman
Enkleia thorelii (Lecomte) Nevling | Enkleia | thorelii | NA | NA | TRUE | (Lecomte) Nevling | Enkleia thorelii | 1961 | accepted | Enkleia thorelii (Lecomte) Nevling | species | LCVF | NA | Enkleia thoreli
Enkleia zippeliana Hallier f. | Enkleia | zippeliana | NA | NA | TRUE | Hallier f. | Enkleia zippeliana | NA | synonym | Enkleia paniculata (Merr.) Hallier f. | species | LCVF | Enkleia paniculata | Enklei
Enkylia digyna Griff. | Enkylia | digyna | NA | NA | TRUE | Griff. | Enkylia digyna | 1845 | synonym | Gynostemma pentaphyllum (Thunb.) Makino | species | LCVF | Gynostemma pentaphyllum | Gynostemma | penta
Enkylia trigyna Griff. | Enkylia | trigyna | NA | NA | TRUE | Griff. | Enkylia trigyna | 1845 | synonym | Gynostemma pentaphyllum (Thunb.) Makino | species | LCVF | Gynostemma pentaphyllum | Gynostemma | per
Enlychnia floresii F. Ritter & Y. Itô | Enlychnia | floresii | NA | NA | TRUE | F. Ritter & Y. Itô | Enlychnia floresii | 1981 | unresolved | unresolved | unresolved | LCVF | NA | NA | NA | NA | NA | NA | NA
Enlychnia longispina F. Ritter & Y. Itô | Enlychnia | longispina | NA | NA | TRUE | F. Ritter & Y. Itô | Enlychnia longispina | 1981 | unresolved | unresolved | unresolved | LCVF | NA | NA | NA | NA | NA | NA | NA
Enneadynamis polonorum Bubani | Enneadynamis | polonorum | NA | NA | TRUE | Bubani | Enneadynamis polonorum | 1901 | synonym | Parnassia palustris L. | species | LCVF | Parnassia palustris | Parnassia | palu
Ennealophus amazonicus N. E. Br. | Ennealophus | amazonicus | NA | NA | TRUE | N. E. Br. | Ennealophus amazonicus | 1909 | synonym | Ennealophus foliosus (Kunth) Ravenna | species | LCVF | Ennealophus folios
Ennealophus boliviensis (Baker) Ravenna | Ennealophus | boliviensis | NA | NA | TRUE | (Baker) Ravenna | Ennealophus boliviensis | 1977 | accepted | Ennealophus boliviensis (Baker) Ravenna | species | L
Ennealophus euryandrus (Gris.) Ravenna | Ennealophus | euryandrus | NA | NA | TRUE | (Gris.) Ravenna | Ennealophus euryandrus | NA | synonym | Ennealophus euryandrus (Griseb.) Ravenna | species | GBIF | Z
Ennealophus euryandrus (Griseb.) Ravenna | Ennealophus | euryandrus | NA | NA | TRUE | (Griseb.) Ravenna | Ennealophus euryandrus | 1973 | accepted | Ennealophus euryandrus (Griseb.) Ravenna | species | L
Ennealophus euryandrus subsp. euryandrus | Ennealophus | euryandrus | subsp. | euryandrus | NA | NA | Ennealophus euryandrus subsp. euryandrus | NA | autonym | Ennealophus euryandrus subsp. euryand
Ennealophus euryandrus subsp. oxydandrus Ravenna | Ennealophus | euryandrus | subsp. | oxydandrus | TRUE | Ravenna | Ennealophus euryandrus subsp. oxydandrus | 2005 | accepted | Ennealophus euryand
Ennealophus fimbriatus Ravenna | Ennealophus | fimbriatus | NA | NA | TRUE | Ravenna | Ennealophus fimbriatus | 1983 | accepted | Ennealophus fimbriatus Ravenna | species | LCVF | NA | Ennealophus fimbriatu
Ennealophus foliosus (Kunth) Ravenna | Ennealophus | foliosus | NA | NA | TRUE | (Kunth) Ravenna | Ennealophus foliosus | 1977 | accepted | Ennealophus foliosus (Kunth) Ravenna | species | LCVF | NA | Enne
Ennealophus foliosus subsp. amazonicus (N. E. Br.) Ravenna | Ennealophus | foliosus | subsp. | amazonicus | TRUE | (N. E. Br.) Ravenna | Ennealophus foliosus subsp. amazonicus | NA | synonym | Ennealo
Ennealophus foliosus subsp. amazonicus (N. E. Br.) Ravenna | Ennealophus | foliosus | subsp. | amazonicus | TRUE | (N. E. Br.) Ravenna | Ennealophus foliosus subsp. amazonicus | 1977 | accepted | Ennealo
Ennealophus foliosus subsp. foliosus | Ennealophus | foliosus | subsp. | foliosus | NA | NA | Ennealophus foliosus subsp. foliosus | NA | autonym | Ennealophus foliosus subsp. foliosus | subspecies | G
Ennealophus foliosus subsp. spruceanus (Baker) Ravenna | Ennealophus | foliosus | subsp. | spruceanus | TRUE | (Baker) Ravenna | Ennealophus foliosus subsp. spruceanus | 1977 | synonym | Ennealophus
Ennealophus simplex (Ravenna) Roitman & J. A. Castillo | Ennealophus | simplex | NA | NA | TRUE | (Ravenna) Roitman & J. A. Castillo | Ennealophus simplex | 2007 | accepted | Ennealophus simplex (Raven
Ennealophus tucumanensis Huaylla | Ennealophus | tucumanensis | NA | NA | TRUE | Huaylla | Ennealophus tucumanensis | 2015 | accepted | Ennealophus tucumanensis Huaylla | species | LCVF | NA | Ennealophus t
Enneapogon abyssinicus (Hochst.) Rendl | Enneapogon | abyssinicus | NA | NA | TRUE | (Hochst.) Rendl | Enneapogon abyssinicus | 1899 | synonym | Enneapogon cenchroides (Licht. ex Schult.) C. E. Hubb
Enneapogon arenicola (Domin) N. T. Burb. | Enneapogon | arenicola | NA | NA | TRUE | (Domin) N. T. Burb. | Enneapogon arenicola | 1941 | synonym | Enneapogon nigricans (R. Br.) P. Beauv. | species | LCVF | EN
Enneapogon asperatus C. E. Hubb. | Enneapogon | asperatus | NA | NA | TRUE | C. E. Hubb. | Enneapogon asperatus | 1937 | accepted | Enneapogon asperatus C. E. Hubb. | species | LCVF | NA | Enneapogon asperatus | Er
Enneapogon avenaceus (Lindl.) C. E. Hubb. | Enneapogon | avenaceus | NA | NA | TRUE | (Lindl.) C. E. Hubb. | Enneapogon avenaceus | 1934 | accepted | Enneapogon avenaceus (Lindl.) C. E. Hubb. | species | LCVF
Enneapogon benguellensis Rendl | Enneapogon | benguellensis | NA | NA | TRUE | Rendl | Enneapogon benguellensis | 1899 | synonym | Enneapogon scaber Lehm. | species | LCVF | Enneapogon scaber | Enneapog
Enneapogon borealis (Griseb.) Honda | Enneapogon | borealis | NA | NA | TRUE | (Griseb.) Honda | Enneapogon borealis | 1936 | synonym | Enneapogon desvauxii P. Beauv. | species | LCVF | Enneapogon desvau
Enneapogon brachystachyus (Jaub. & Spach) Stapf | Enneapogon | brachystachyus | NA | NA | TRUE | (Jaub. & Spach) Stapf | Enneapogon brachystachyus | 1900 | synonym | Enneapogon desvauxii P. Beauv. | sy
Enneapogon brachystachyus var. macrantherus Stapf | Enneapogon | brachystachyus | var. | macrantherus | TRUE | Stapf | Enneapogon brachystachyus var. macrantherus | 1900 | synonym | Enneapogon desv
Enneapogon caerulescens (Gaudich.) N. T. Burb. | Enneapogon | caerulescens | NA | NA | TRUE | (Gaudich.) N. T. Burb. | Enneapogon caerulescens | 1941 | accepted | Enneapogon caerulescens (Gaudich.) N. T.
Enneapogon caerulescens var. caerulescens | Enneapogon | caerulescens | var. | caerulescens | NA | NA | Enneapogon caerulescens var. caerulescens | NA | autonym | Enneapogon caerulescens var. caerul
Enneapogon caerulescens var. occidentalis Kakudidi | Enneapogon | caerulescens | var. | occidentalis | TRUE | Kakudidi | Enneapogon caerulescens var. occidentalis | 1989 | synonym | Enneapogon caeri
Enneapogon cenchroides (Licht. ex Schult.) C. E. Hubb. | Enneapogon | cenchroides | NA | NA | TRUE | (Licht. ex Schult.) C. E. Hubb. | Enneapogon cenchroides | NA | accepted | Enneapogon cenchroides (Li
Enneapogon cenchroides (Licht.) C. E. Hubb. | Enneapogon | cenchroides | NA | NA | TRUE | (Licht.) C. E. Hubb. | Enneapogon cenchroides | 1934 | synonym | Enneapogon cenchroides (Licht. ex Schult.) C. E.

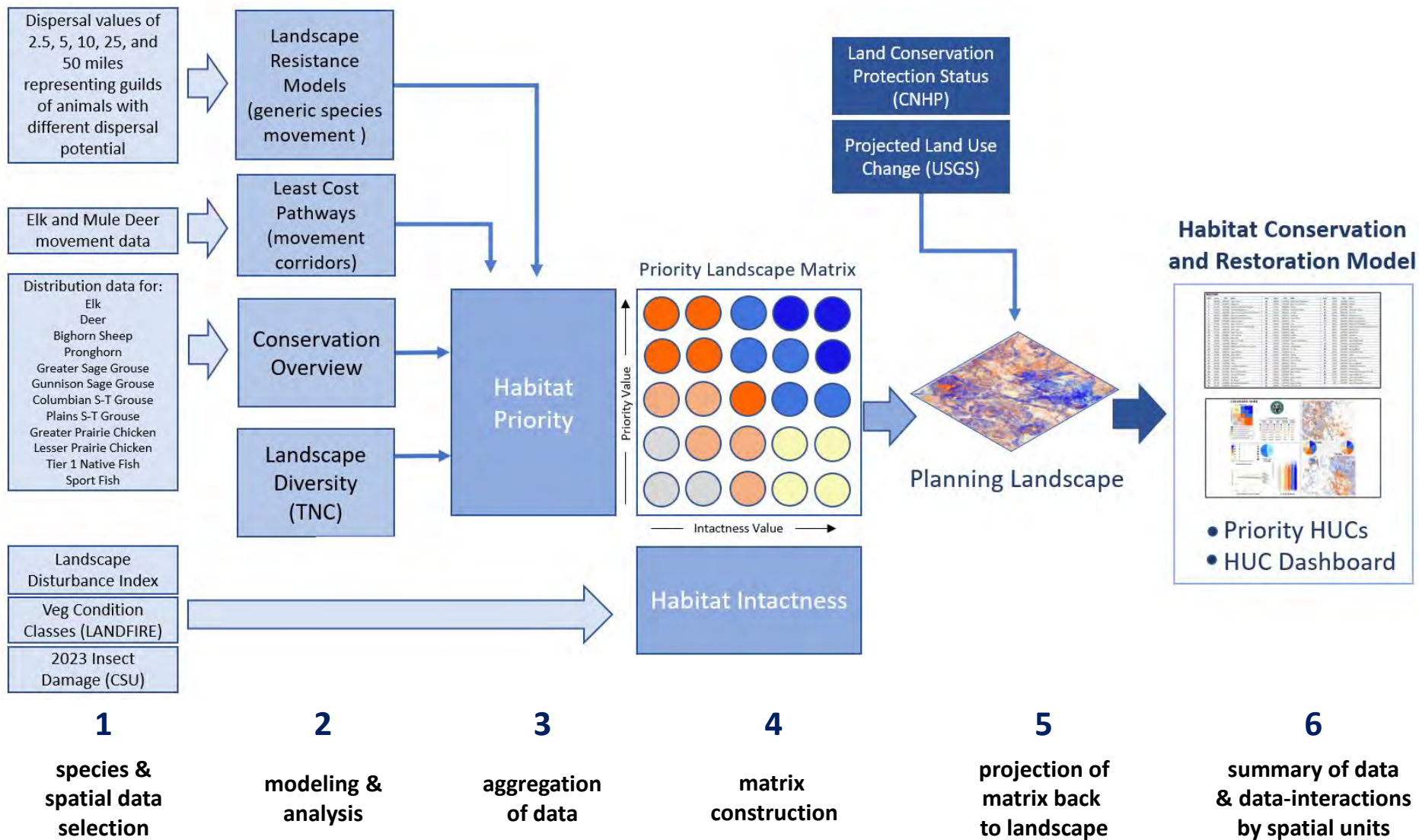
1,831,687 vascular plant names and name variants

[illegible]

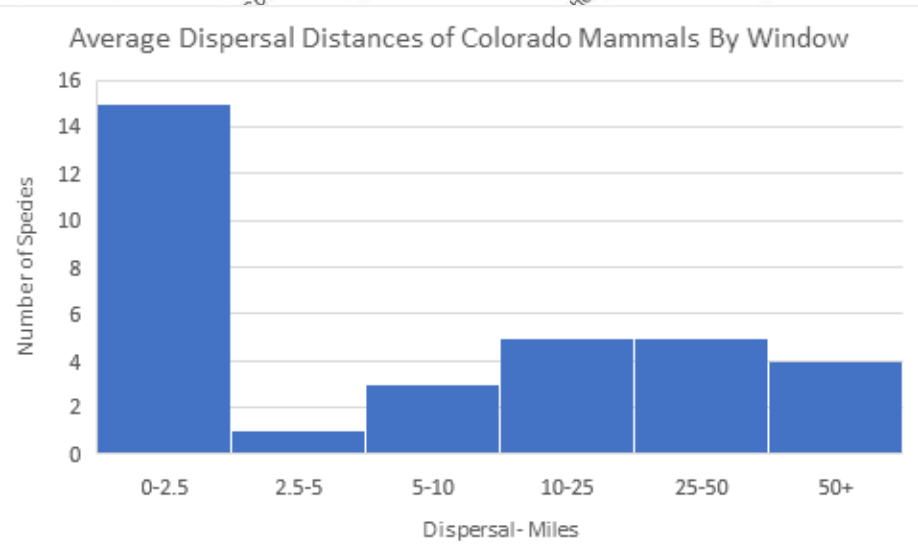
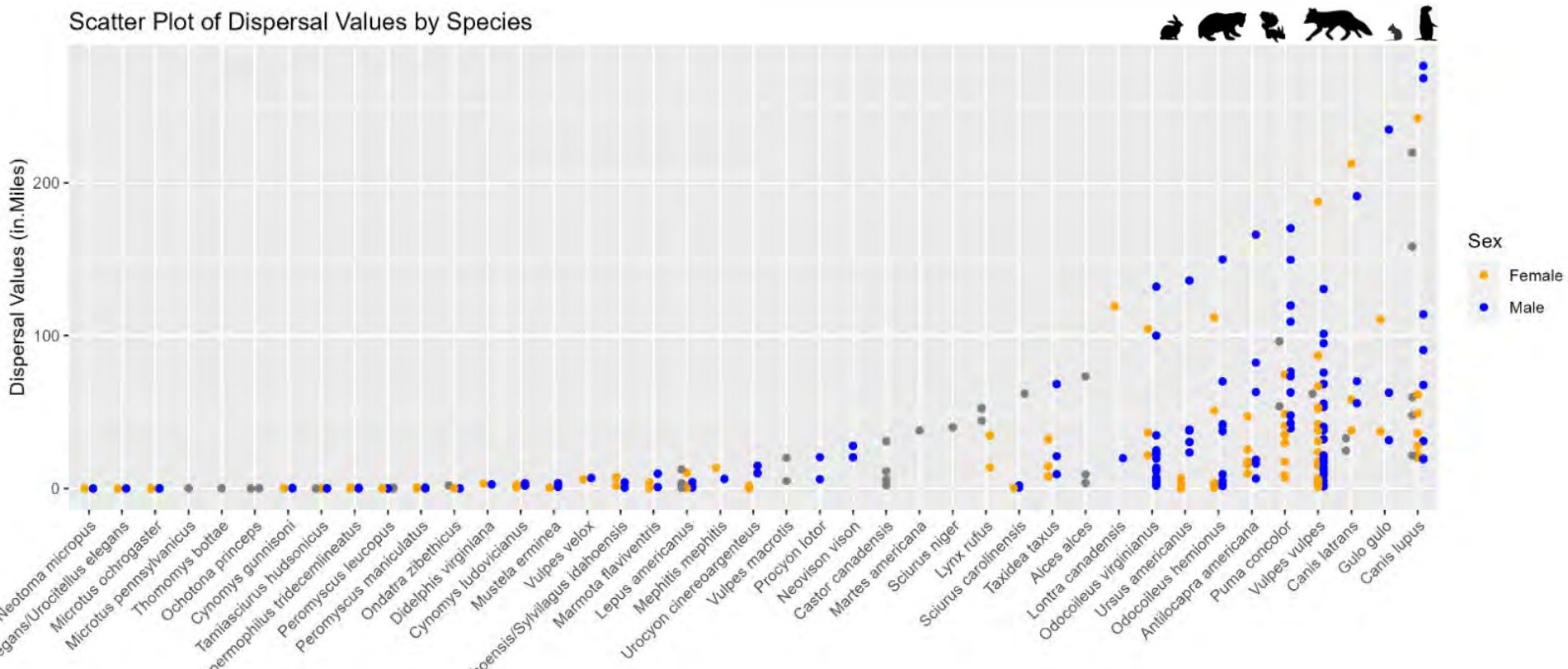


COLORADO HABITAT CONSERVATION AND CONNECTIVITY PLAN DASHBOARD



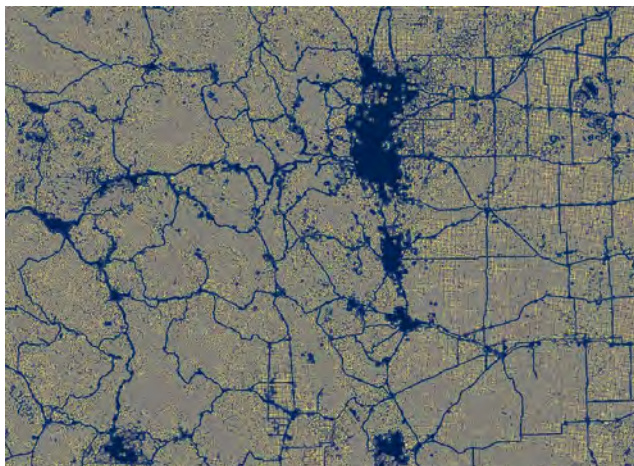


Ominscape is agnostic of individual species traits and represents generalized movement potential

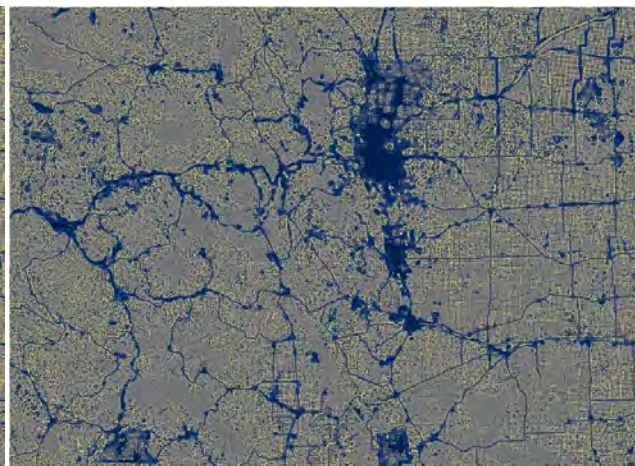


Movement for Colorado mammals skews local

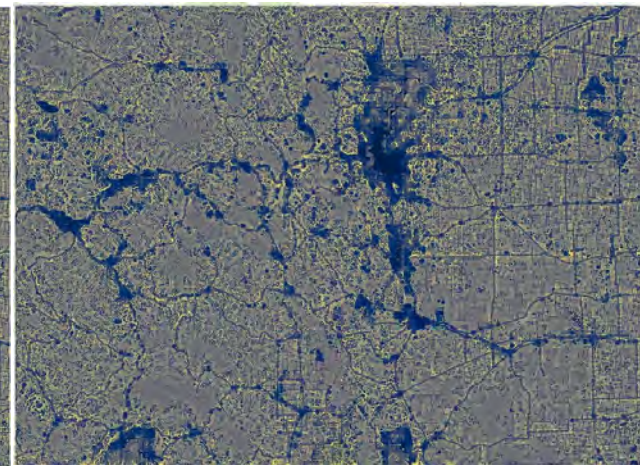
0.25, 5, 10, 25, and 50 mile radii were used in this study



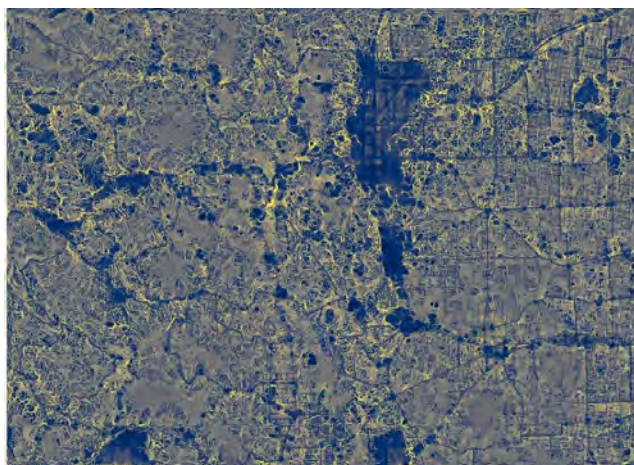
2.5 miles



5 miles



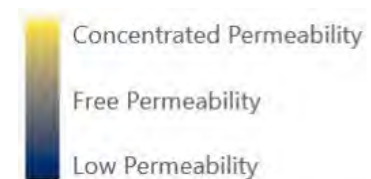
10 miles

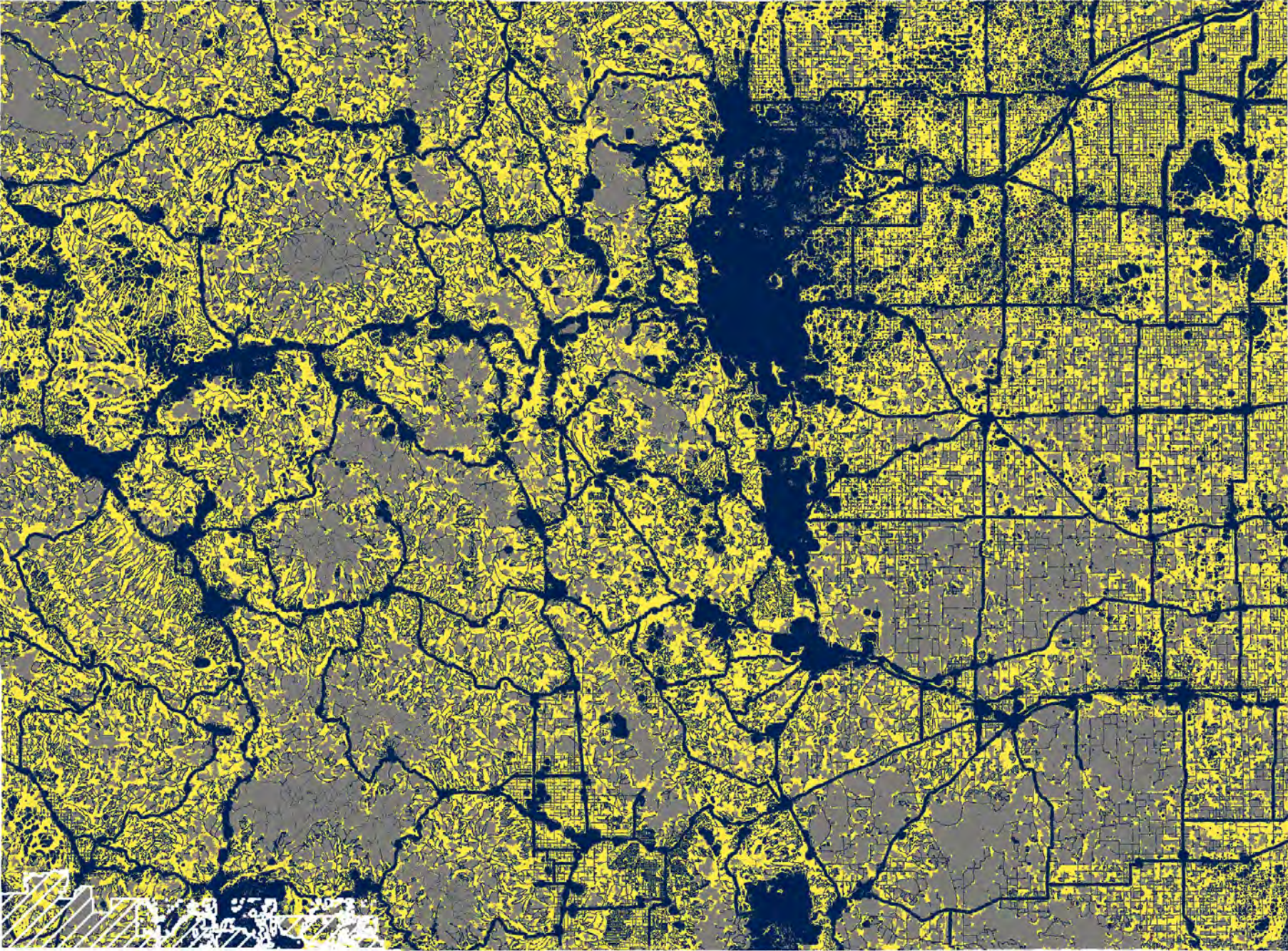


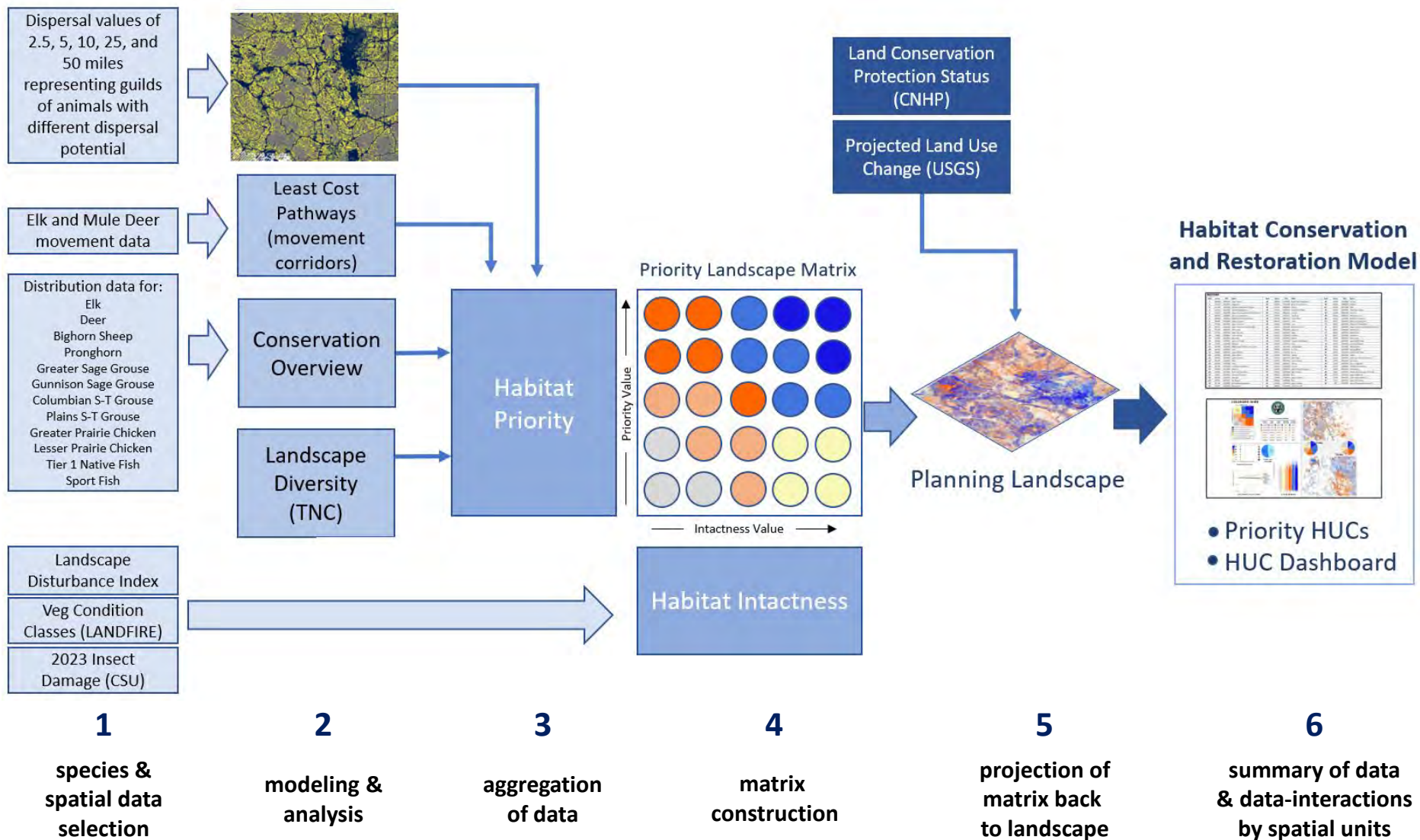
25 miles

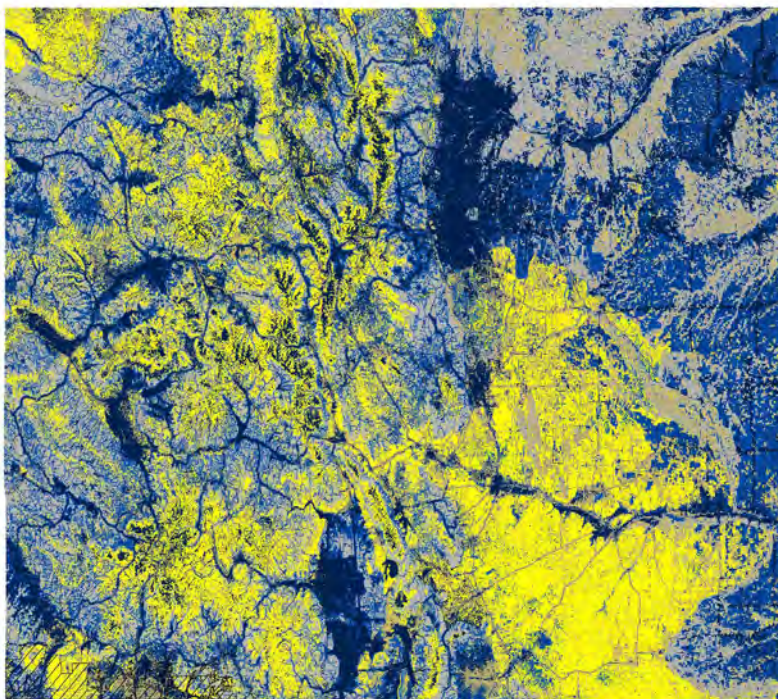


50 miles

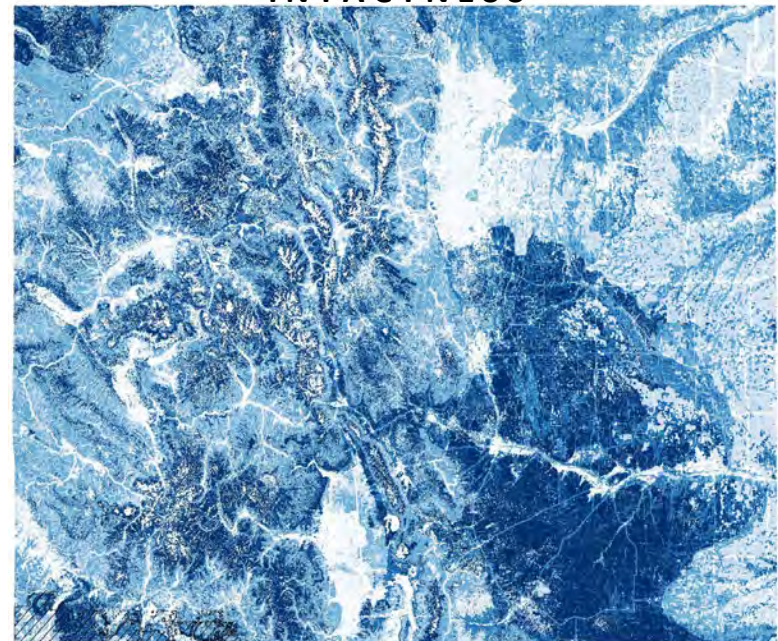
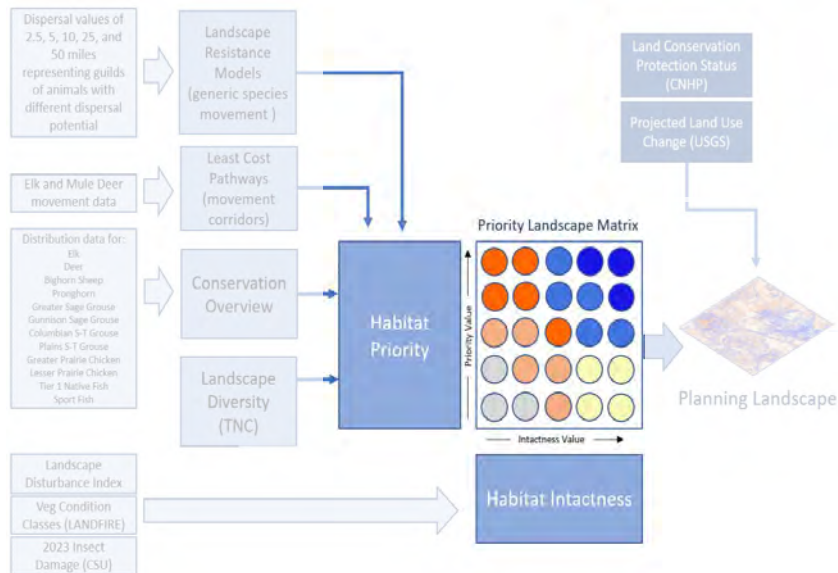
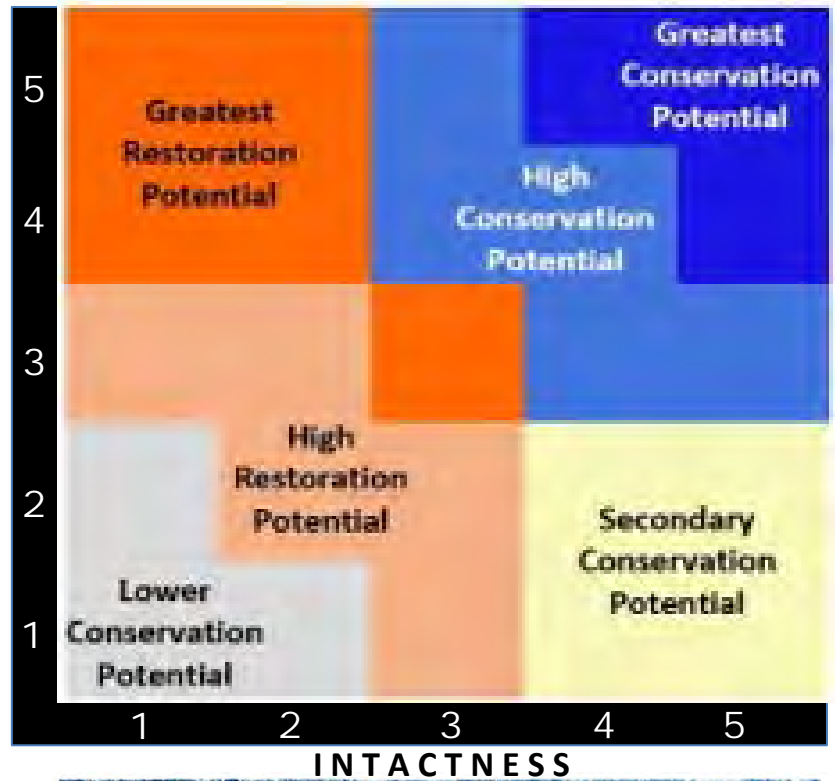


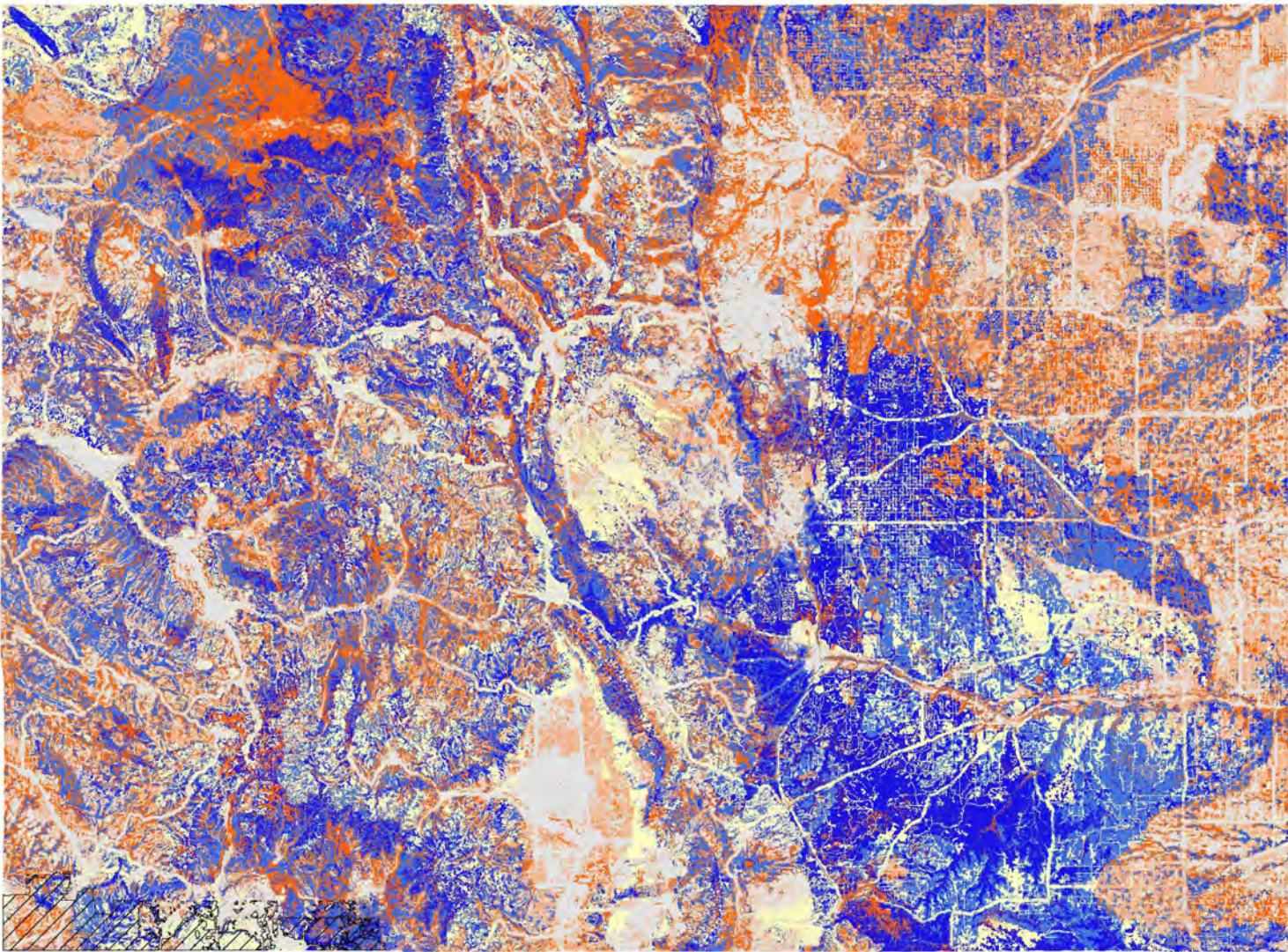




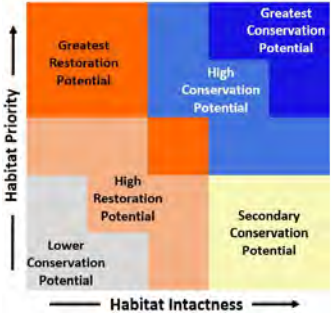


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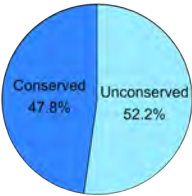
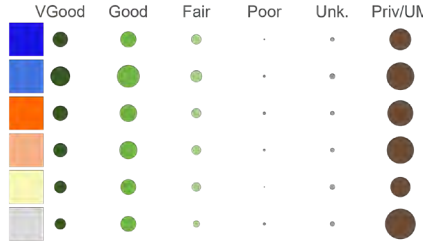
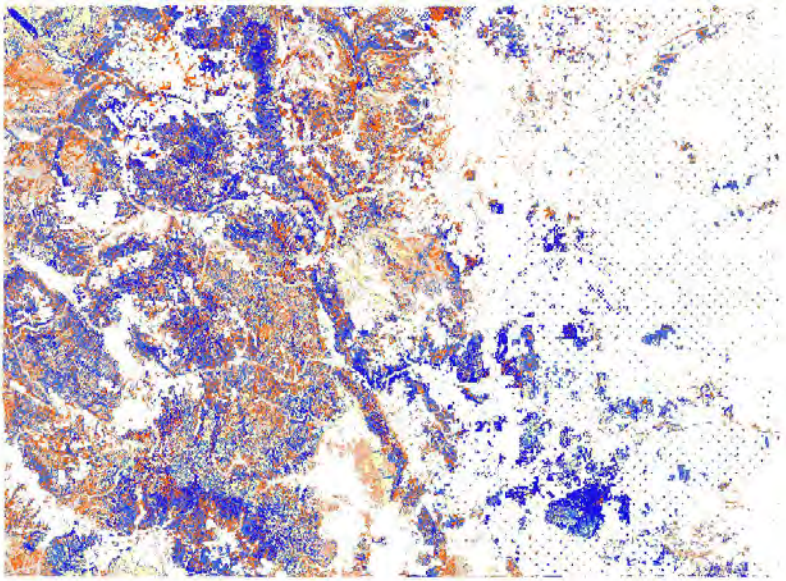


COLORADO



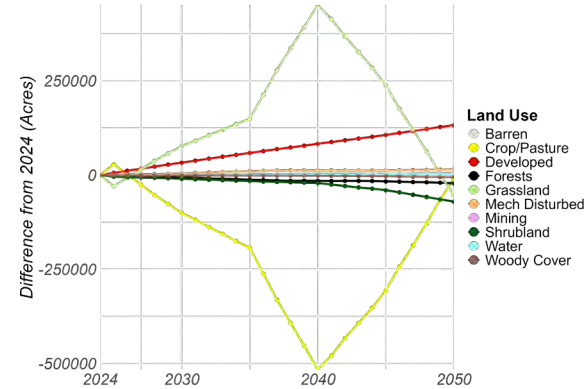
Landscape Summary Statistics

	Total Area (acres)	Conserved (acres)	Unconserved (acres)	Aggregation (out of 100)	Cohesion (out of 100)
	8,898,313	4,766,693	4,021,916	84.56	98.70
	15,827,261	8,933,664	6,696,712	84.45	98.65
	11,081,396	5,189,446	5,717,744	82.21	99.05
	11,592,918	5,044,816	6,393,939	83.42	98.33
	7,467,712	3,960,664	3,407,847	82.65	98.80
	11,751,269	3,408,508	8,138,105	89.41	99.91
	66,618,872	31,303,794	34,376,266	84.45	98.91

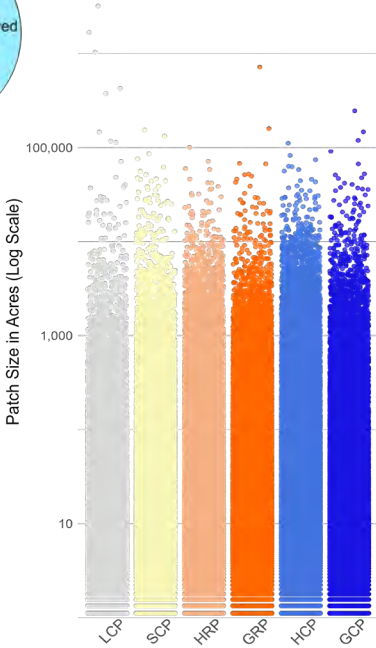


Current Land Ownership

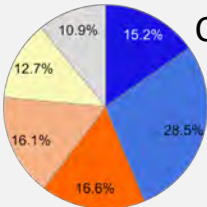
Current Land Protection



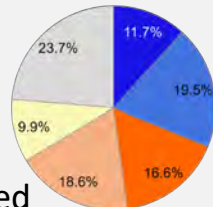
Future Predicted Land Use Change



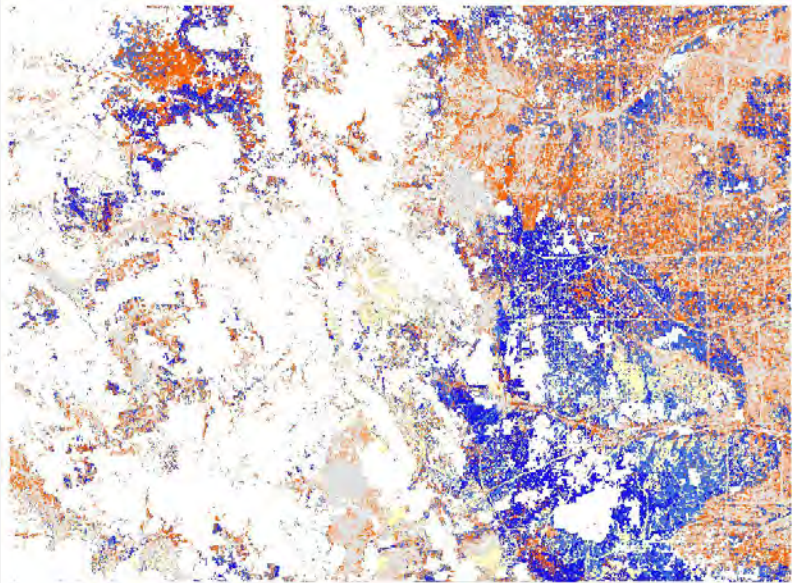
Patch Size Distribution



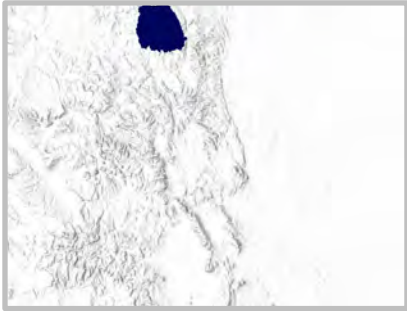
Conserved



Unconserved



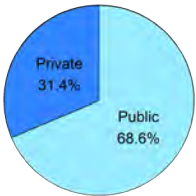
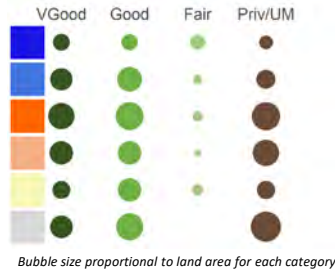
NORTH PLATTE HEADWATERS 10180001



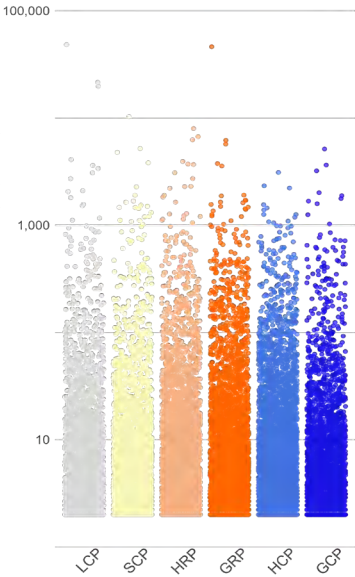
Protect (Potential Land Conservation) rank: 43/92
Defend (Maintain Land Conservation) rank: 28/92
Restore (Habitat Restoration) rank: 32/92
Combined Rank: 30/92 (tied)



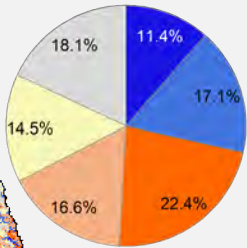
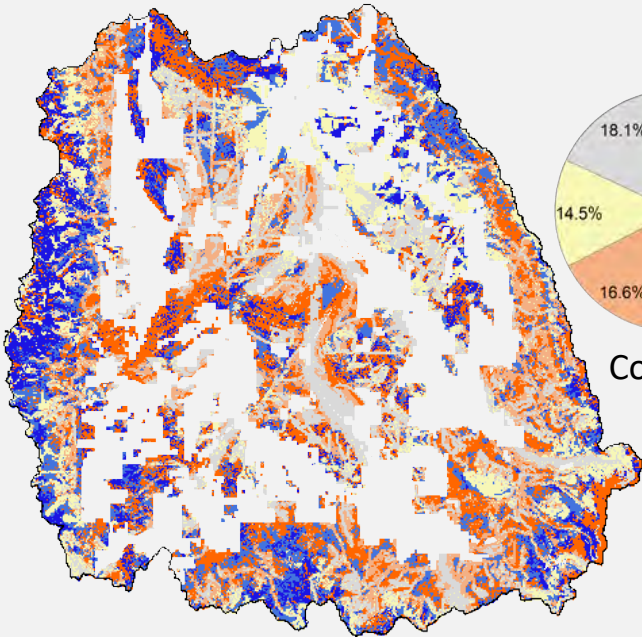
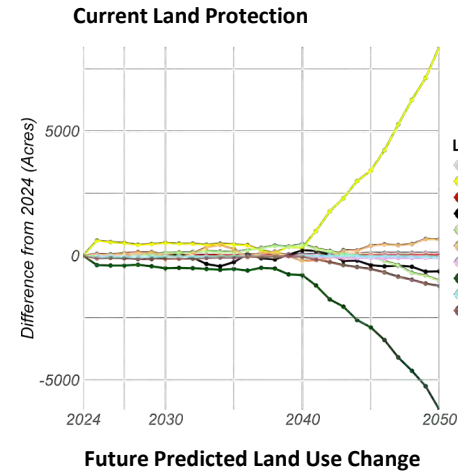
Landscape Summary Statistics				
Total Area (acres)	Conserved (acres)	Unconserved (acres)	Aggregation (out of 100)	Cohesion (out of 100)
88,117	70,189	16,317	75.58	96.84
139,820	105,314	32,076	74.12	95.71
209,852	137,993	67,396	81.70	98.96
163,930	102,233	59,048	78.29	97.82
119,921	89,073	28,627	76.54	97.15
191,075	111,282	75,832	85.07	99.26
912,718	616,086	279,298	78.55	97.62



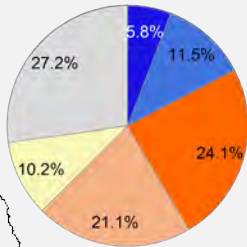
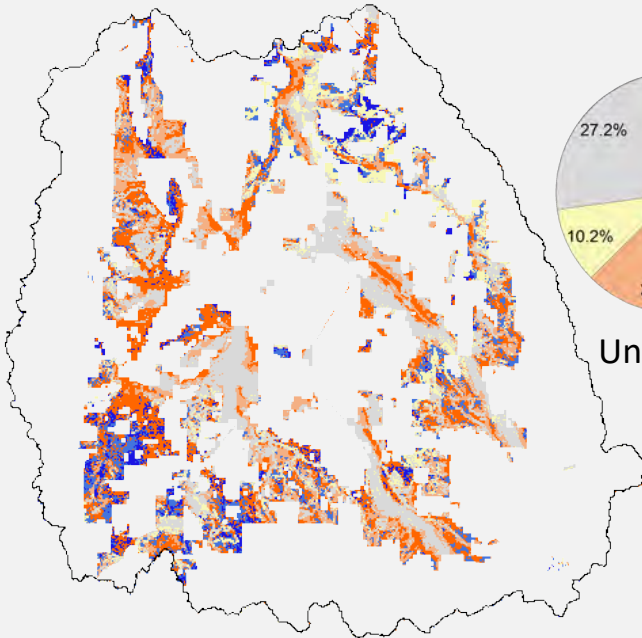
Current Land Ownership



Patch Size Distribution

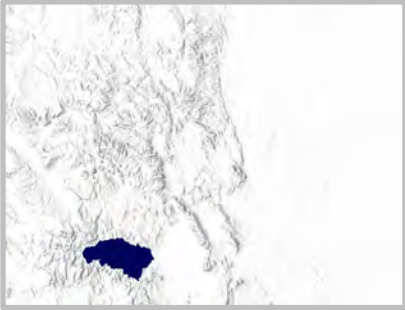


Conserved



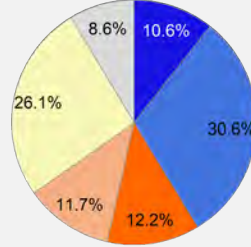
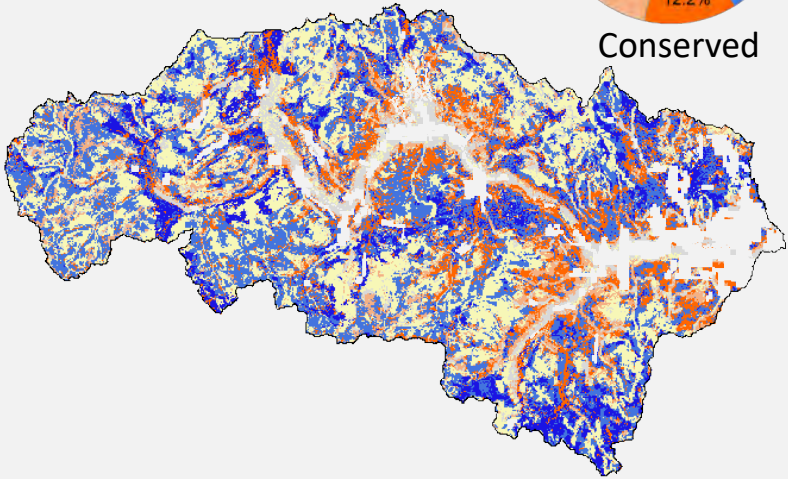
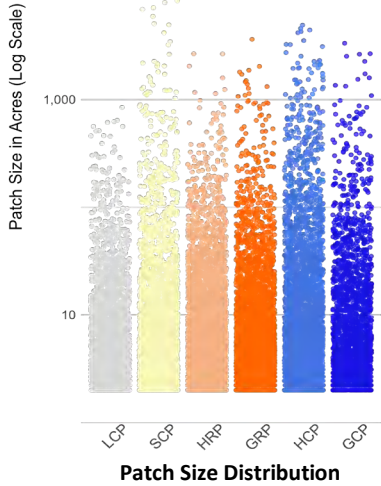
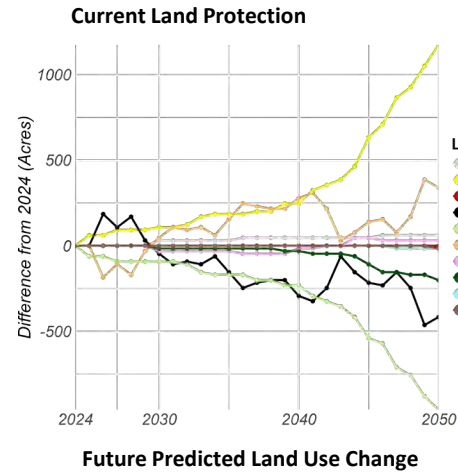
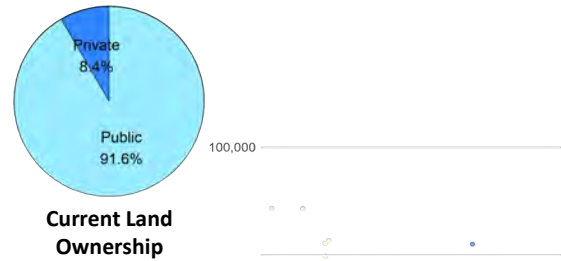
Unconserved

RIO GRANDE HEADWATERS 13010001

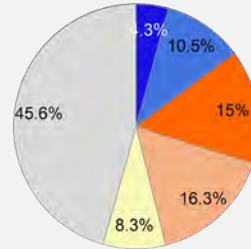
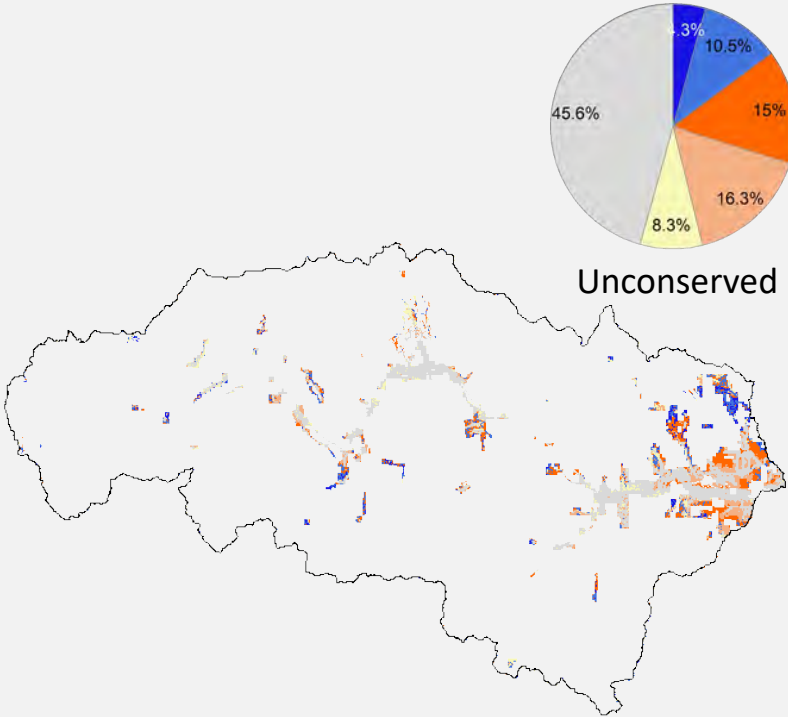


Protect (Potential Land Conservation) rank:	72/92
Defend (Maintain Land Conservation) rank:	13/92
Restore (Habitat Restoration) rank:	30/92
Combined Rank:	40/92

Landscape Summary Statistics				
Total Area (acres)	Conserved (acres)	Unconserved (acres)	Aggregation (out of 100)	Cohesion (out of 100)
89,520	85,630	3,000	79.03	96.78
255,789	246,599	7,431	80.15	98.10
110,610	98,494	10,591	73.21	95.94
107,535	94,490	11,495	71.45	94.59
217,402	210,041	5,822	84.94	98.54
104,411	69,465	32,186	81.18	99.01
885,269	804,722	70,528	78.33	97.16

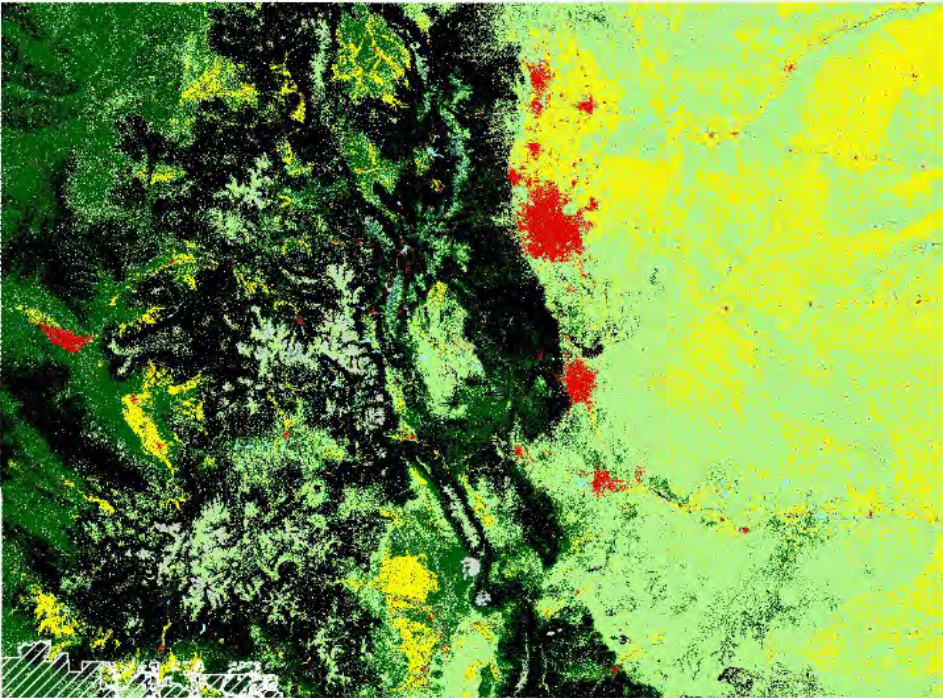


Conserved

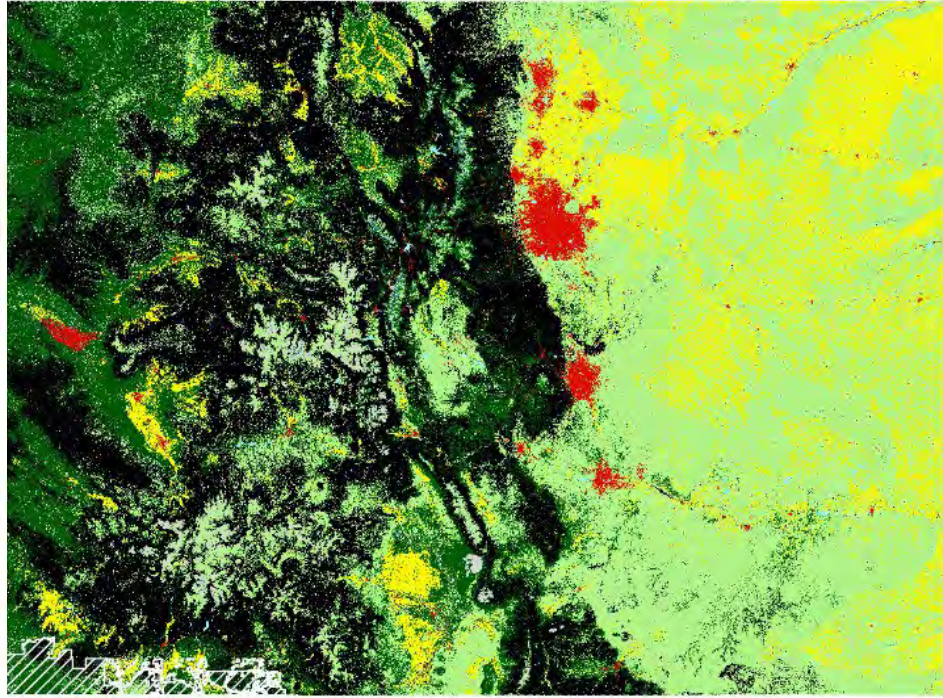


Unconserved

2024



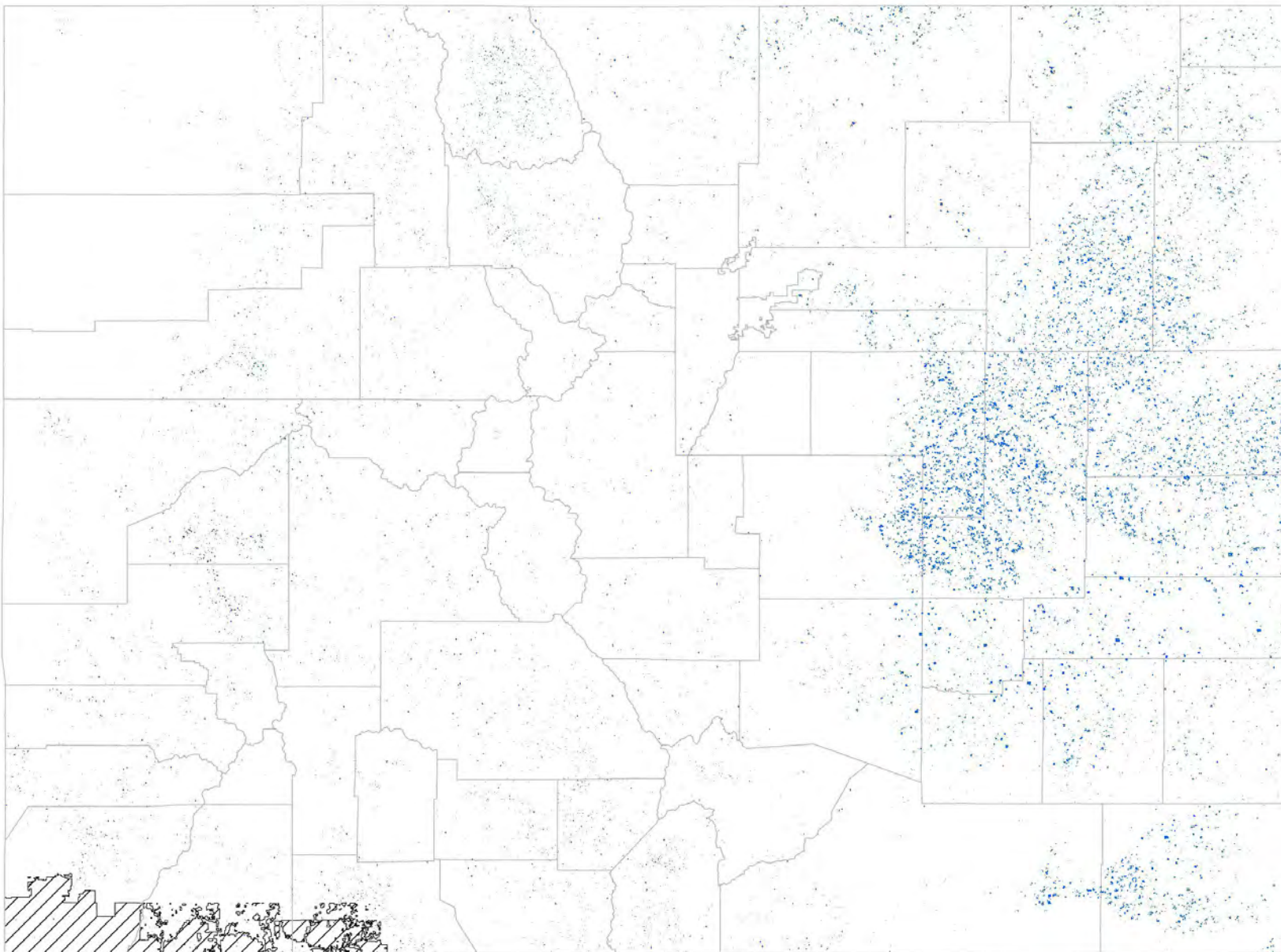
2050



USGS FORE-SCE Model

- Uses the **IPCC SRES “A2”** scenario (socioeconomic impacts on anthropogenic land use incorporating demographics, energy use, agricultural economics, and other socioeconomic considerations).
- A2 is the “worst case” scenario, with continued economic development and temperature increases 2.0-5.4 degrees C.
- The FORE-SCE model “known” landscape data from 1992-2005 using the A2 scenario, then projects 2006-2100.
- Predicts future land use and land cover.

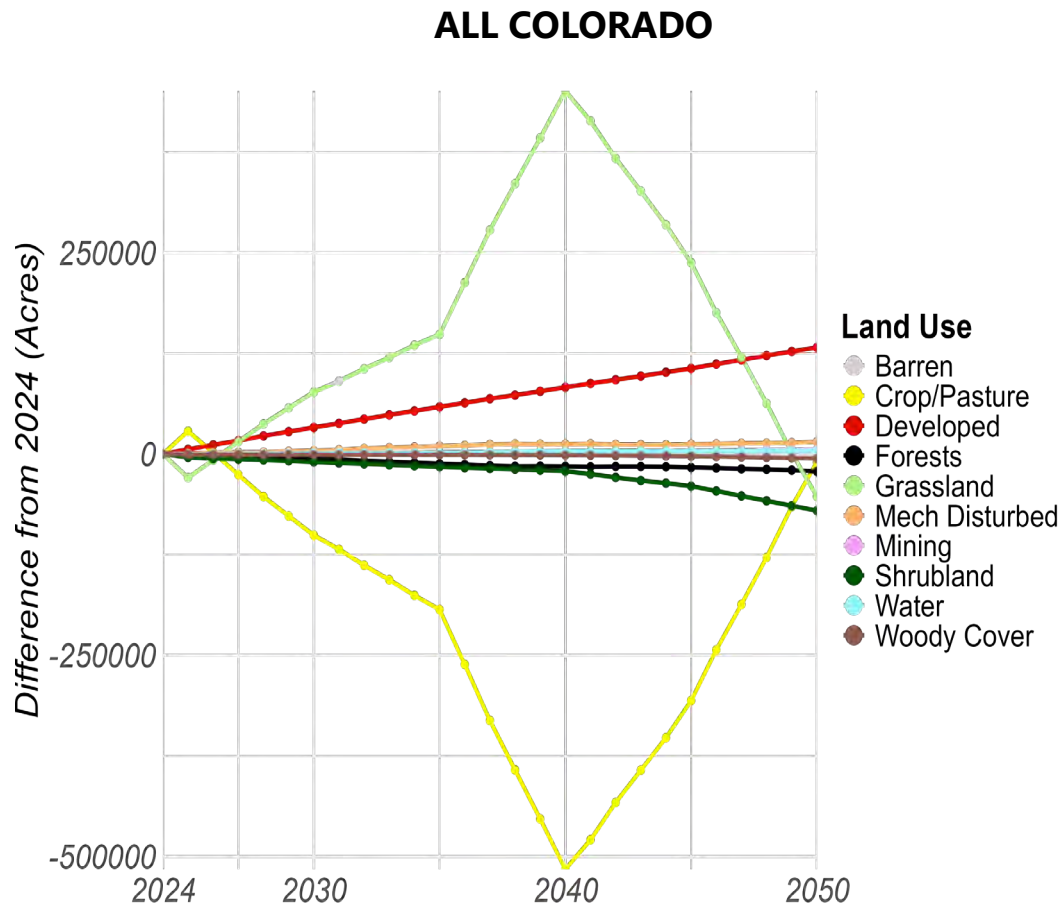




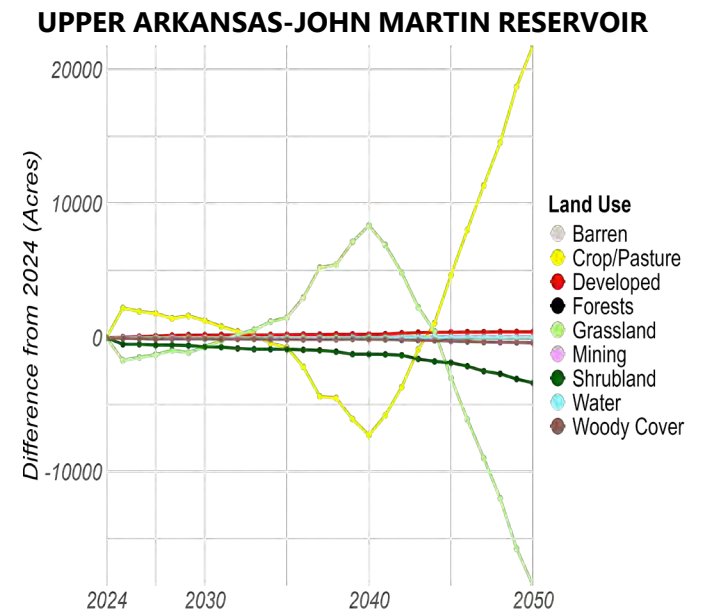
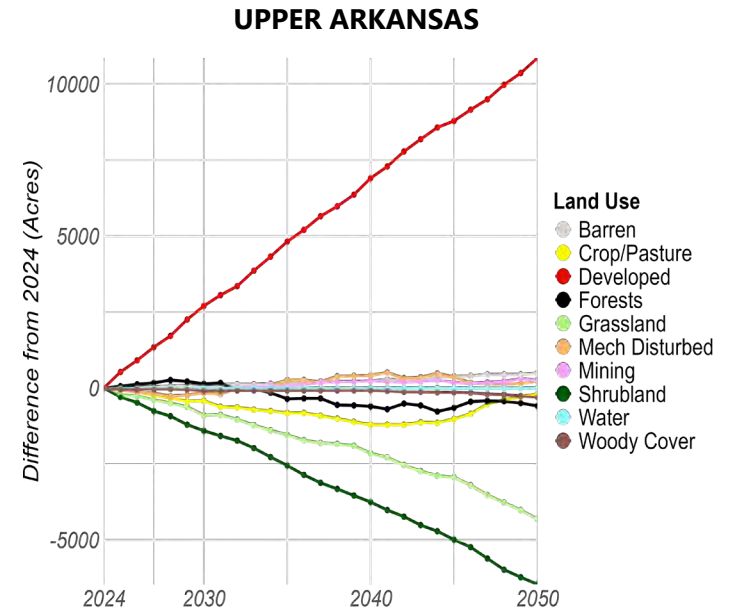
**Projected Land Use Change 2024 vs 2050
Under USGS A2 Model**

■ Land Use Change
 Tribal Land

— Colorado County Boundaries



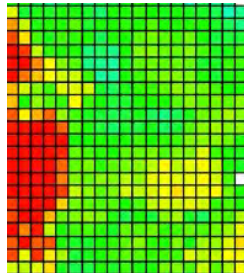
PROJECTED LAND USE CHANGE



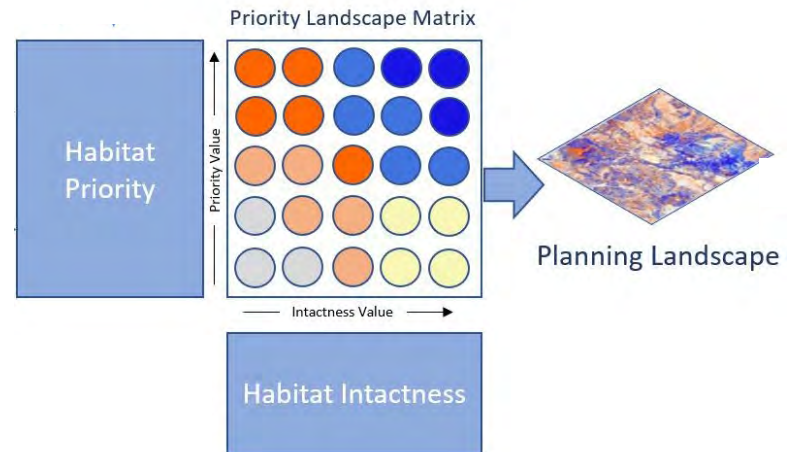
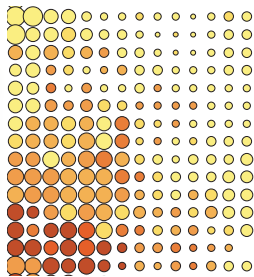
Future Hopes & Dreams



Colorado
Species
Diversity



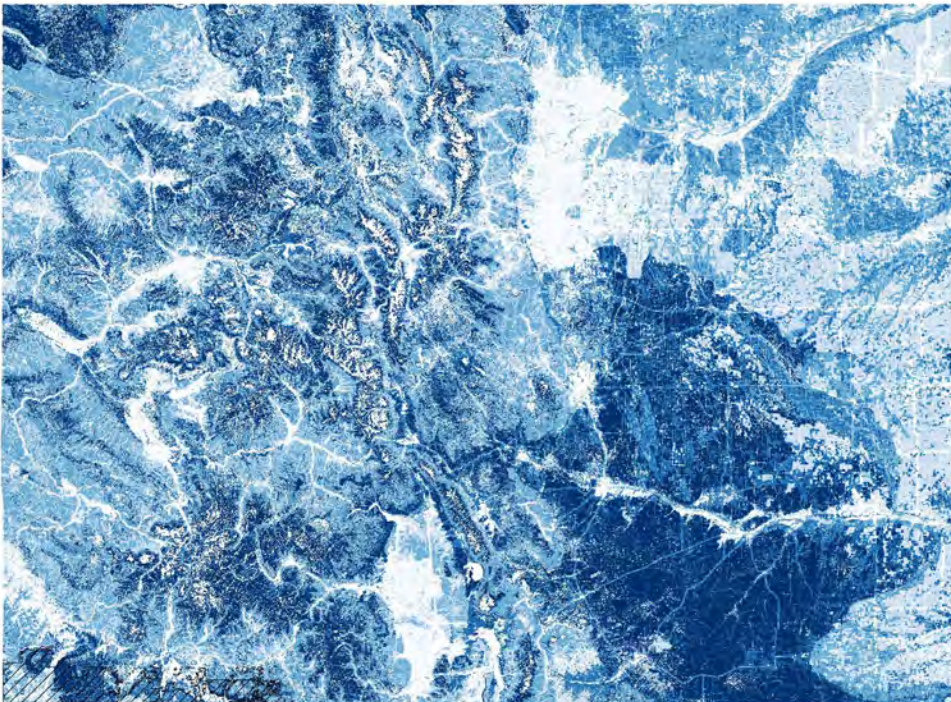
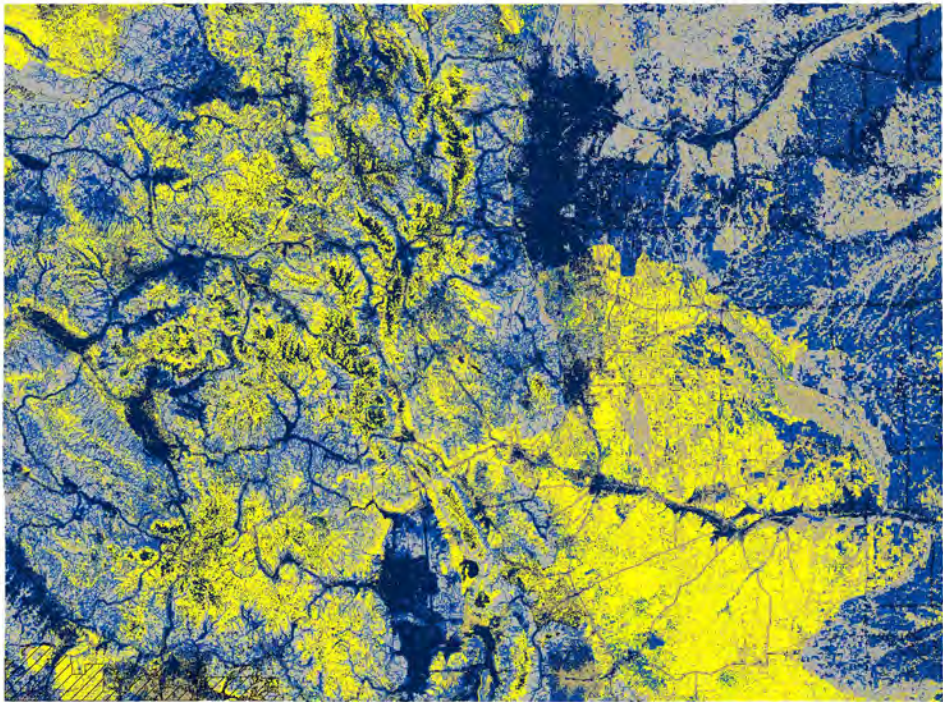
Colorado
Genetic
Diversity



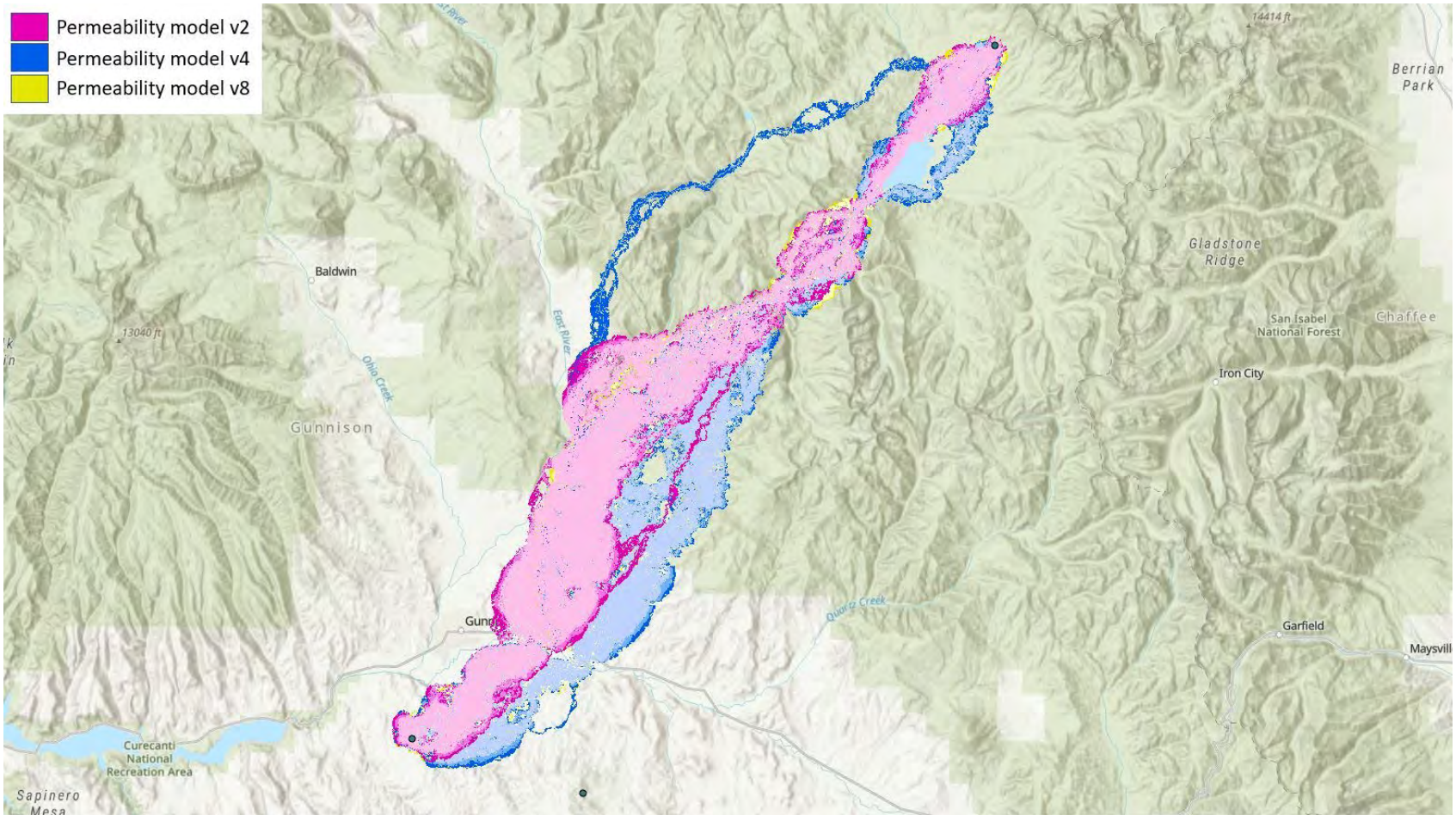




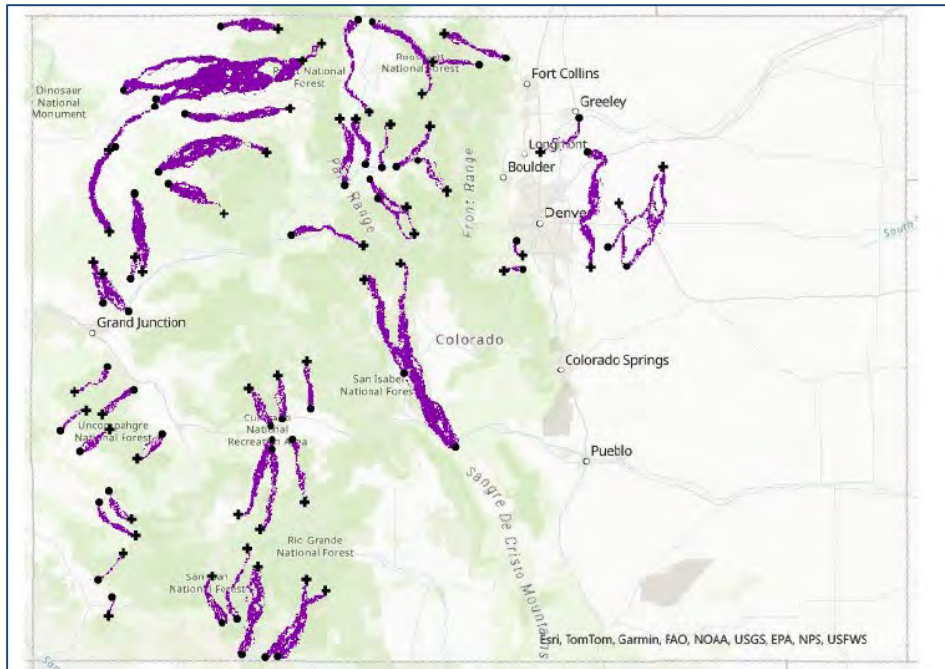




- Ran a set of “known” corridor paths using a subset of the candidate models
- Tested the overlap of those models with the know movement paths



The final model was then used to extrapolate between “unknown” start and end points



Mule Deer



Elk